Prepared for **Nevada Environmental Response Trust**

Project Number **1690011200-034**

Prepared by **Ramboll Emeryville, California**

Date **November 27, 2019**

PHASE 6 GROUNDWATER FLOW AND TRANSPORT MODEL

NEVADA ENVIRONMENTAL RESPONSE TRUST SITE HENDERSON, NEVADA

Project No **21-38800C** Ramboll 2200 Powell Street Suite 700 Emeryville, CA 94608 USA T +1 510 655 7400 F +1 510 655 9517 www.ramboll.com

Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site (Former Tronox LLC Site) **Henderson, Nevada**

Nevada Environmental Response Trust (NERT) Representative Certification

I certify that this document and all attachments submitted to the Division were prepared at the request of, or under the direction or supervision of NERT. Based on my own involvement and/or my inquiry of the person or persons who manage the system(s) or those directly responsible for gathering the information or preparing the document, or the immediate supervisor of such person(s), the information submitted and provided herein is, to the best of my knowledge and belief, true, accurate, and complete in all material respects.

Office of the Nevada Environmental Response Trust

Le Petomane XXVII, Inc., not individually, but solely in its representative capacity as the Nevada Environmental Response Trust Trustee Not Individually, but Solely

Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site (Former Tronox LLC Site) Henderson, Nevada

Responsible Certified Environmental Manager (CEM) for this Project

I hereby certify that I am responsible for the services described in this document and for the preparation of this document. The services described in this document have been provided in a manner consistent with the current standards of the profession and, to the best of my knowledge, comply with all applicable federal, state and local statutes, regulations and ordinances.

Solarpe

John M. Pekala, PG Principal

November 27, 2019

Date

Certified Environmental Manager Ramboll CEM Certificate Number: 2347 CEM Expiration Date: September 20, 2020

The following individuals provided input to this document:

John M. Pekala, PG Allan J. DeLorme, PE Christopher M. Stubbs, PhD, PE Alka Singhal, PhD Pejman Rasouli, PhD Lisa A. Taylor, MESM

Ruhui Zhao, MS Mia Sosa, MS Jon Hunt, PhD Katie Zheng, MS Annika Deurlington

Date **November 27, 2019** Prepared by **Ramboll** Description **Phase 6 Groundwater Flow and Transport Model**

Project No **1690011200-034**

Ramboll 2200 Powell Street Suite 700 Emeryville, CA 94608 USA T +1 510 655 7400 F +1 510 655 9517 www.ramboll.com

CONTENTS

LIST OF TABLES

- Table 1 Estimated Perchlorate Mass Discharge to Las Vegas Wash
- Table 2a Groundwater Discharge to Las Vegas Wash Based on Chemistry and Flow Data
- Table 2b Estimated Water Balance in Las Vegas Wash
- Table 2c Groundwater Extraction Rates
- Table 2d Conceptual Groundwater Balance Summary
- Table 2e Evaluation of Uncertainty in Conceptual Boundary Inflows
- Table 3 Model Layers
- Table 4 Estimated versus Calibrated Hydraulic Conductivity and Storage Properties
- Table 5 Weir Specifications
- Table 6 Simulated Groundwater Inflows and Outflows
- Table 7 Calibration Statistics for Flow Model
- Table 8 Sensitivity Analysis of Flow Model Results
- Table 9 Conceptual Perchlorate Mass Removal Summary
- Table 10 Transport Model Parameters
- Table 11 Calibration Statistics for Transport Model
- Table 12 Sensitivity Analysis of Transport Model Results
- Table 13 Modeled Perchlorate Mass Removal Summary

LIST OF FIGURES

- Figure 1-1 Model Location Map
- Figure 2-1 Model Area: Geology and Groundwater Flow
- Figure 2-2 Las Vegas Wash and Other Hydrology Features
- Figure 2-3 Model Area Features: Remediation System
- Figure 2-4 Timeline for Major Remediation Events in the Model Domain
- Figure 3-1 Streamflow at Las Vegas Wasteway Gage Station
- Figure 3-2 Streamflow at Duck Creek Gage Station
- Figure 3-3 Streamflow at Duck Creek Confluence Gage Station
- Figure 3-4 Streamflow at Pabco Road Gage Station
- Figure 3-5 Streamflow at Bostic Road Gage Station
- Figure 3-6 Streamflow at Homestead Gage Station
- Figure 3-7 Streamflow at Three Kids Gage Station
- Figure 3-8 Schematic Representing Various Las Vegas Wash Water Balance Components
- Figure 3-9 Perchlorate Mass Loading at the Las Vegas Wash
- Figure 3-10 AECOM Transducers Data versus Stream Stages
- Figure 3-11 Remediation Pumping in the Model Domain (2014-2018)
- Figure 3-12 Non-Remediation Pumping in the Model Domain
- Figure 3-13 Model Area Features: Unlined Ponds
- Figure 3-14 Model Area Features: Storm Drains
- Figure 3-15 Model Area Features: Sub-Drains
- Figure 3-16 Conceptual Boundary Inflows
- Figure 4-1 E-W Geological Model Cross-Sections
- Figure 4-2 N-S Geological Model Cross-Sections
- Figure 4-3 Alluvium Thickness
- Figure 4-4 Coarse-grained Upper Muddy Creek Formation Thickness
- Figure 4-5 Fine-grained Upper Muddy Creek Formation Thickness
- Figure 4-6 Coarse-grained Upper Muddy Creek Formation Depositional Series

APPENDICES

ACRONYMS AND ABBREVIATIONS

1. INTRODUCTION

On behalf of the Nevada Environmental Response Trust (the Trust), Ramboll US Corporation (Ramboll) has prepared this "Phase 6 Groundwater Flow and Transport Model" report for the Nevada Environmental Response Trust Site (the NERT Site or Site), located in Clark County, Nevada. The Phase 6 Model is a comprehensive groundwater model of the southeastern portion of the Las Vegas Groundwater Basin that was developed to support the Remedial Investigation and Feasibility Study (RI/FS) of the NERT Site. Figure 1-1 shows the model extent.

The focus of the Phase 6 Model development was to update the previous Phase 5 Model to incorporate new Phase 2 RI, Phase 3 RI, and Downgradient Investigation data and to expand the capabilities of the model to simulate perchlorate contaminant transport. In addition to RI and Downgradient Investigation data, the model incorporates available subsurface investigation data collected by all parties, including the Trust and its consultants, other companies within the Black Mountain Industrial (BMI) Complex, American Pacific Corporation (AMPAC)/Endeavour LLC, and public agencies (e.g., Southern Nevada Water Authority [SNWA], United States Geological Survey [USGS], Nevada Department of Environmental Pollution [NDEP], and Bureau of Reclamation [BOR]). Recent data collected during the Downgradient Investigation and future data to be collected during the Operable Unit 3 (OU-3) RI will be incorporated into the next version of the model. In addition, the next model version will include the capability to simulate transport of other major contaminants of potential concern (COPCs).

Throughout the remainder of this report the Phase 6 Model will be referred to simply as "the model", except where there is a need to differentiate between various versions.

1.1 Model Objectives

The model is a key tool for supporting the NERT RI/FS, the ongoing optimization of the existing groundwater extraction and treatment system (GWETS), and other related activities. The transport model will be used to predict the effect of remedial measures on mass discharge to Las Vegas Wash ("the Wash") and remediation timeframes. The transport model is also being used to calculate the expanded performance metrics, as described in the NDEP-approved RI Study Area Mass Estimate and Expanded Performance Metrics Technical Approach Technical Memorandum (Ramboll Environ 2017), dated October 5, 2017, and to evaluate the effectiveness of the existing GWETS and potential future remedial action alternatives in the forthcoming FS.

The model is designed to include all potential source areas of perchlorate and other NERT COPCs, so that the model can be used to better understand the relative contribution of different source areas to discharges into the Las Vegas Wash and impacts to other receptors. In addition to providing a tool for synthesizing multiple dat.a sources and further developing the conceptual site model, the model can be used for predicting the future impacts to receptors like the Las Vegas Wash under alternative remediation strategies and technologies. This provides a way of systematically comparing the relative effectiveness and timeframe of these alternatives in meeting remedial action objectives.

1.2 Phase 6 Model Refinements

The Phase 6 Model was calibrated to simulate transient groundwater flow conditions between 2014 and 2018. This period was selected because it has a robust concentration data set available to use for calibration of the transport model.

In addition to including the new data in the model, the model has been refined to address NDEP comments on the Phase 5 Model (NDEP 2017). Below is the list of refinements included in the Phase 6 Model:

- The potential impact of density effects on the flow system behavior has been evaluated and is described in Appendix A.
- A complete review of any permitted groundwater pumping (other than the remediation pumping) was completed.
- A detailed assessment of the uncertainty in the boundary recharge fluxes was done and is described in this report.
- The Weston Hills and Chimera Golf course area sub-drains were added to the model.
- The model representation of the Las Vegas Wash was updated to represent conditions at the beginning of 2018.
- The alluvium thickness in the model has been refined based on the borings installed as part of the RI Phase 2, RI Phase 3, and other investigations conducted since 2016.
- The conceptual groundwater discharge to the Wash has been updated based on perchlorate loading at all USGS stream gage stations within the model domain.
- The quarterly conceptual water balance for the model domain has been refined based on additional data collected as part of the RI.
- The model was refined and calibrated based on the additional vertical gradient data collected as part of the RI.
- The vertical conductivities were updated in the model to better calibrate groundwater heads in artesian wells in the model domain.
- The evaporation rates applied to the Wash and the Henderson Bird Viewing Preserve ponds have been updated.
- The southern, western, and the northern model boundaries were updated from general head boundary conditions to specified flow boundary conditions.
- The groundwater model has been re-calibrated to match average quarterly water level targets from 2014-2018.
- The capability to simulate perchlorate transport was added to the model. The transport model was calibrated to perchlorate mass flux to the Wash, perchlorate mass removal from the remediation well fields, and perchlorate discharge to residential subdrains.

1.3 Report Organization

The report is organized as follows. The model area background and a conceptual water balance for the modeled period is described in Section 2.0 and Section 3.0, respectively. The refinements in the 3D geological model is described in Section 4.0. A description of the groundwater flow model development is provided in Section 5.0, with model

calibration described in Section 6.0. Section 7.0 describes the transport modeling approach, Section 8.0 the transport model development, and Section 9.0 the transport model calibration and results. Conclusions are provided in Section 10.0 and references in Section 11.0.

An evaluation of the effects of density on the flow system behavior is provided in Appendix A. The annual distribution of phreatophytes and recharge zones within the model domain are shown on figures presented in Appendix B and Appendix C, respectively. A summary of hydraulic testing results is presented in Appendix D. Measured groundwater elevations and vertical gradient data are presented in Appendix E and Appendix F, respectively. Observed versus simulated head residuals for each model target location during the simulation period are provided in Appendix G. Appendix H has a table with measured perchlorate concentrations at model targets, while Appendix I presents figures showing measured and simulated perchlorate concentrations at each target location. Finally, Appendix J has the model input files. Appendices E through J are provided electronically.

2. MODEL AREA BACKGROUND

The model area is located within the Las Vegas Valley in the southern region of Clark County, Nevada. Las Vegas Valley is surrounded by a set of mountain ranges, including the Spring Mountains to the west, the Sheep Range and Las Vegas Range to the north, the Frenchman Mountains and Sunrise Mountains to the east, and the River Mountains and McCullough Mountains to the south (Figure 1-1). The most significant stream in the valley is the Las Vegas Wash, which flows generally from west to east before discharging into Lake Mead. The climate in the area varies from semi-arid in the mountains to arid in the lowlands. Rainfall averages about 5 inches per year and occurs in storms of high intensity and short duration that often lead to floods. Potential evaporation in the area is significant and can be higher than 80 inches per year in the lower portion of the valley (UNLV 2003).

This section provides background on the model area. The first two subsections describe the geology and hydrogeology of the model area, based on the Tech Memo of RI Data Evaluation (Ramboll Environ 2016c). This is followed by a description of Las Vegas Wash hydrology and environmental remediation activities within the model area. Finally, a summary of previous modeling work is presented.

2.1 Geology

The mountain ranges bounding the east, north, and west sides of the valley consist primarily of Paleozoic and Mesozoic sedimentary rocks (limestones, sandstones, siltstones, and fanglomerates), whereas the mountains on the south and southeast consist primarily of Tertiary volcanic rocks (basalts, rhyolites, andesites, and related rocks) that overlie Precambrian metamorphic and granitic rocks (ENSR 2007).

In the Las Vegas Valley, eroded Tertiary and Quaternary sedimentary and volcanic rocks comprise the unconsolidated basin deposits, which can be up to 13,000 feet (ft) thick (ENSR 2007). The valley floor consists of fluvial, paludal (swamp), playa, and lacustrine deposits surrounded by more steeply sloping alluvial fan aprons derived from erosion of the surrounding mountains (Figure 2-1). Generally, the deposits grade finer with increasing distance from their source and with decreasing elevation. The structure within the Quaternary and Tertiary-aged basin fill is characterized by a series of generally north-south trending fault scarps. Within the model domain, three main hydrogeologic units are present: alluvium, Transitional Muddy Creek Formation, and Upper Muddy Creek Formation. In addition, in the northeast corner of the model domain, the Horse Spring Formation is present near Las Vegas Wash. These geological units are described below:

Alluvium

The uppermost unit is composed of Quaternary alluvial deposits that slope north toward Las Vegas Wash. The alluvium consists of a reddish-brown heterogeneous mixture of well-graded sand and gravel with lesser amounts of silt, clay, and caliche (Figure 2-1). Clasts within the alluvium are primarily composed of volcanic material. Boulders and cobbles are common. Due to the mode of deposition, no distinct beds or units are continuous over the model area.

A major feature of the alluvial deposits is the stream-deposited sands and gravels that were laid down within paleochannels eroded into the surface of the Muddy Creek Formation during infrequent flood runoff periods. These deposits vary in thickness and are narrow and generally linear. These generally uniform sand and gravel deposits exhibit higher permeability than the adjacent, well-graded deposits. In general, these paleochannels trend northeastward. Recent deposits near the Wash are coarse sand and gravels, described as "wash gravels" in the remainder of the report. These deposits are highly permeable and are present beneath and on both sides of the Wash (Figure 2-1).

The thickness of the alluvial deposits ranges from less than 1 foot to more than 50 ft beneath the Site. Soil types identified in on-site soil borings include poorly sorted gravel, silty gravel, poorly sorted sand, well sorted sand, and silty sand. The thickness of the alluvium, as well as the top of the underlying Muddy Creek Formation, was mapped to locate these paleochannels*.*

Transitional (or reworked) Muddy Creek Formation

Where present, Transitional Muddy Creek Formation (xMCf) is encountered at the base of the alluvium. The Transitional Muddy Creek Formation consists of reworked sediments derived from the Muddy Creek Formation, which is described below. Therefore, the xMCf appears similar to the Muddy Creek Formation, but it consists of reworked, less consolidated and indurated sediments.

Upper Muddy Creek Formation

The Upper Muddy Creek Formation (UMCf) of Pleistocene age occurs in the Las Vegas Valley as valley-fill deposits that are coarse-grained near mountain fronts and become progressively finer-grained toward the center of the valley. Where encountered beneath the Site, the Muddy Creek Formation is composed of at least two thicker units of finegrained sediments of clay and silt (the first and second fine-grained facies) interbedded with at least two thinner units of coarse-grained sediments of sand, silt, and gravel (the first and second coarse-grained facies). Except for the southernmost 1,000 ft adjacent to Lake Mead Parkway, the first fine-grained facies (UMCf-fg1) separates the first coarsegrained facies (UMCf-cg1) from the overlying Quaternary alluvium at the Site. Within the southern 1,000 ft of the Site, the Muddy Creek Formation's UMCf-fg1 pinches out along a roughly west-northwesterly trending line. South of this line, the UMCf-cg1 directly underlies the Quaternary alluvium. Other coarse-grained units are found within the UMCf to the east and west of the NERT Site, described further in Section 4.0.

Locally, the UMCf represents deposition in an alluvial apron environment from the Spring Mountains to the west, grading into fluvial, paludal (swamp), playa, and lacustrine environments further out into the valley center. On the Site, the Muddy Creek does not crop out but instead subcrops beneath a veneer of Quaternary alluvium. Since the Site is located closer to the mountains, the upper portion of the UMCf-fg1 unit tends to have zones of sandy silt/silty fine sand as well as a greater number of thin, discontinuous layers of silty sand than in the downgradient plume area, which is farther from the mountains and more toward the interior of the depositional basin.

Horse Spring Formation

In the model area near the Three Kids weir, the Horse Spring Formation and a subunit called the Thumb Member underlie the Las Vegas Wash alluvium. The Horse Spring Formation is a valley fill deposit, consisting of a yellowish-brown dolomitic limestone interbedded with siltstone (GES 2003). The Thumb Member of the Horse Spring Formation consists of reddish brown, interbedded mudstone, and calcareous sandstone with variable amounts of gypsum.

2.2 Hydrogeology

The depth of groundwater varies from near the ground surface for wells near the Wash to about 100 ft below the ground surface (ft bgs) for wells in the southernmost portion of the NERT Site. The groundwater flow direction in the model area is generally to the north (Figure 2-1). This generally uniform flow pattern may be modified locally by subsurface alluvial channels cut into the underlying UMCf (paleochannels), areas of localized recharge from artificial ponds and trenches, and groundwater extraction from remediation system well fields.

NDEP has defined three water-bearing zones (WBZs) that are of interest in the BMI Complex: the Shallow WBZ, which is defined by the first occurrence of groundwater in either the alluvium, xMCf, or the UMCf where the xMCf is missing, is unconfined to partially confined, and is considered the "water table aquifer" (within the depth interval from 0 to 90 ft bgs); the Middle WBZ, which extends from approximately 90 to 300 ft bgs; and the Deep WBZ, which is defined as the contiguous WBZ that is generally encountered between 300 to 400 ft bgs (NDEP 2009). Environmental investigations at the Site have primarily focused on the Shallow WBZ, although recent investigations have included a number of Middle WBZ wells to improve vertical delineation of hydrogeology and chemical constituent distribution.

2.3 Hydrology

Las Vegas Valley is drained by Las Vegas Wash, a 12-mile-long channel which flows into Lake Mead. Accounting for less than 2 percent of the water in Lake Mead, the water flowing through the Wash consists of urban runoff, shallow groundwater, storm water and treated wastewater from the Clark County Sanitation District, the City of Henderson, and the City of Las Vegas. Prior to the development of Las Vegas Valley, the Wash was an ephemeral stream. Currently, the flow is perennial and composed almost entirely of effluent from wastewater treatment plants. This increase in flow caused significant erosion and downcutting of the Wash channel. A program of weir construction was conducted by the Las Vegas Wash Coordinating Committee to address erosion and downcutting. The Pabco Road Weir and part of the Historic Weir were constructed in early 2000. Twelve more weirs have been constructed since 2000 within the model domain. The Sunrise Mountain Weir and the expansion of Historic Lateral Weir, the final weirs planned under the construction program, were completed in 2018 (LVWCC 2019). The locations of weirs within the model area are shown on Figure 2-2.

2.4 Removal Actions, RI/FS-Related Activities, and Remediation by Other Parties

This section presents a summary of removal actions and RI/FS-related activities conducted by NERT and other parties within the model area. Major activities include the operation of groundwater pump and treat systems, installation of slurry walls, and reinjection of treated groundwater. The locations of these activities are presented on Figure 2-3, and an historical timeline is shown on Figure 2-4.

2.4.1 NERT Well Fields

The NERT Groundwater Extraction and Treatment System (GWETS) consists of three well fields that are shown on Figure 2-3 and described below:

Interceptor Well Field

In September 1986, Kerr-McGee Chemical LLC (Kerr-McGee) entered into a Consent Order with Nevada Division of Environmental Protection (NDEP) to undertake groundwater remediation. Installation of the Interceptor Well Field (IWF), processing equipment and a recharge trench was completed during 1987 to remediate chromium impacts to the alluvium. The system has been operating since that time, with semiannual reports detailing the removal action effort regularly submitted to NDEP.

In 1994, Kerr-McGee completed the Underground Injection Control (UIC) permitting process for the treated groundwater injection trench, associated with this removal action effort. The permit was subsequently modified in December 1998 to allow for injection of Lake Mead water, rather than the treated groundwater (Kerr-McGee 2001). This in turn allowed the treated groundwater to be placed into a holding pond, awaiting further treatment before reinjection or discharge. Beginning in 1998, groundwater recovered and treated by the chromium treatment plant was discharged to an onsite 11-acre pond. This change was instituted to halt recharge of perchlorate-impacted groundwater back to the sub-surface system (Kerr-McGee 2000). Untreated stabilized Lake Mead water was injected through the recharge trenches at or below the rate of groundwater extraction by the interceptor wells. The recharge trenches ceased operation in September 2010. The barrier wall was emplaced downgradient of the IWF in 2001 (ENVIRON 2014b). The locations of IWF pumping wells, barrier wall, and the recharge trenches are shown on Figure 2-3.

AP Area Wells

The AP Area Down and Up Flushing Treatability Study, located approximately 300 ft south of the IWF just west of AP-5, was implemented in 2016 and operated independently of the IWF (Ramboll 2018c). Initial soil flushing and extraction well testing was conducted in October 2016, with continuous operation beginning in November 2016 with three extraction wells, E1-1, E1-2, and E1-3. Flushing and extraction in Plot 2 began in July 2017 with three new extraction wells E2-1, E2-2, and E2-3. Following completion of the treatability study in early 2018, the AP Area extraction wells continued operating as part of the GWETS as requested by NDEP. The locations of AP Area pumping wells are shown on Figure 2-3.

Athens Road Well Field

The Athens Road Well Field (AWF) was initially designed to intercept perchlorate in groundwater downgradient of the IWF and the Site (Figure 2-3). The AWF is approximately 8,200 ft north (downgradient) of the barrier wall and the IWF. The AWF was constructed as a series of 14 groundwater extraction wells screened in the alluvium at seven paired well locations that span approximately 1,200 ft across two alluvial paleochannels located on either side of an UMCf ridge. The AWF was completed in March 2002 and continuous pumping began in mid-October of that year (Ramboll Environ 2016c).

Seep Well Field

The Seep Well Field (SWF) and the seep capture sump.¹, located approximately 4,500 ft north (downgradient) of the AWF near the Las Vegas Wash, are shown on Figure 2-3 and were installed in response to the discovery of perchlorate-impacted groundwater migrating (via an intermittent surface seep) to the Las Vegas Wash. In the late 1990s/early 2000s, the groundwater captured from the seep sump was conveyed to an ion exchange unit for treatment (ENSR 1999).

When pumping began in July 2002, the SWF consisted of three extraction wells (PC-99R2/R3, PC-115R, and PC-116R) situated over the deepest part of the alluvial channel and a seep capture sump designed to capture the intermittent surface seep. Five additional wells (PC-117, PC-118, PC-119, PC-120, and PC-121) were completed in February 2003 and an additional well (PC-133) was completed in December 2004. Presently, the SWF consists of 10 extraction wells—two of which (PC-99R2 and PC-99R3) are connected and operate as one combined well. The wells comprising the SWF are screened across the full thickness of the alluvium and across the deepest portion of an alluvial channel (Ramboll Environ 2016c).

2.4.2 OSSM Well Field

The Olin Chlor Alkali Products, Montrose Chemical Corporation of California, and Stauffer Management Company (collectively referred to herein as "OSSM") have been operating the Groundwater Treatment System (GWTS) at the OSSM site since December 1983. The GWTS is operational under a Consent Order between the State of Nevada, Stauffer, and Montrose for groundwater remediation (Hargis 2008). As shown on Figure 2-3, the OSSM GWTS consists of a series of extraction wells and recharge trenches.

Originally, there were thirteen OSSM extraction wells (A, B, C, P, J, D2, E3, L, F, G, H2, K2, and I). In 2008, well M was replaced by new pumping well M2. In 2009, two more wells, N and O, were installed. Wells A and B have been offline since 2011, and two new extraction wells, Q and R, were installed and have been operational since October 2011 (Hargis 2012).

2.4.3 AMPAC Well Fields

The former Pacific Engineering and Production Company (PEPCON) facility manufactured perchlorates from approximately 1958 until 1988. The original PEPCON plant footprint covered approximately 15 acres, and AMPAC owned more than 300 contiguous acres. Endeavour LLC (Endeavour) was created in 2015 to conduct remediation activities at the facility.

An In-Situ Bioremediation Treatment System (ISB) was operated from 2006 to 2012 [\(AMPAC](file://ENV-EVL-FILE3/Public/Eng/LePetomane/NERT/Modeling/Remediation%20System%20-Other%20Companies/AMPAC/Docs/2012-08-14%20BCA%20SemiAnnual%20Annual%20Performance%20and%20BWPC%20UIC%20Jan-Jun%202012.pdf) 2012). The treatment system was comprised of an array of nine extraction wells (6 AREWs [Athens Road Extraction Wells] and 3 APEWs [Athens Pen Extraction Wells]), a water handling and conditioning plant, and a re-injection area (Figure 2-3). The groundwater was pumped from the extraction wells to the processing plant where it was blended, filtered, and then pumped north to the re-injection area located approximately 7,000 ft downgradient near the Las Vegas Wash (Figure 2-3). The original re-injection system design consisted of six re-injection wells (RIWs). A seventh

ł

¹ The seep capture sump was reportedly last operated in April 2007 and was decommissioned (pump removed and piping blocked) shortly thereafter. Currently, only the seep sump remains, but it was buried during construction of the Sunrise Mountain Weir by SNWA.

larger diameter well (RIW-5A) was added in July 2008, and a deep re-injection trench (DRIT) located just north of the original RIWs was added in October 2009. The ISB system was shut down in June 2012 and transitioned to an above-ground treatment system (AGTS), which began operation in late September 2012.

Five new extraction wells called Auto Mall Extraction Wells (AMEWs) were installed in January 2012 to extract a higher volume of perchlorate contaminated groundwater from near the former AMPAC site source area. The AMEWs extract groundwater from the Middle WBZ at the depth of 180 ft to 230 ft bgs at a combined rate of approximately 350-450 gpm. An underground effluent HDPE pipeline conveys treated water to a point near Las Vegas Wash. The 14 extraction wells (AMEWs, AREWs, and APEWs) are shown on Figure 2-3.

2.4.4 TIMET Well Field

The TIMET remediation system consists of the slurry wall and groundwater extraction and treatment (GEI 2014). The slurry wall, constructed along the TIMET northern site boundary in 2013, is approximately 2,410 ft in length and extends 60 ft bgs. There is a line of 19 extraction wells screened in the alluvium and extending approximately 3 to 5 ft into the UMCf that was installed in October 2014. The wells are generally located 150 to 160 ft apart and approximately 100 ft upgradient of the slurry wall (Figure 2-3). The alignment of extraction wells EWQal-01 to EWQal-05, located at the northern end, was angled somewhat closer to the wall to capture groundwater that could migrate northward along the wall. At the extreme northern end, well EWQal-01 is located approximately 25 ft from the slurry wall. After treatment using air stripping, the treated water is discharged to a series of six injection trenches located on the downgradient side of the slurry wall.

2.4.5 Weir Dewatering

During the construction of the Sunrise Mountain and Historic Lateral Weirs, groundwater was pumped from dewatering trenches between first quarter 2018 through the third quarter 2018. The water was treated at a temporary treatment plant and then discharged to the Wash near the existing NERT Outfall. The dewatering trench locations are shown on Figure 2-3.

2.5 Previous Groundwater Modeling Studies

Previous versions of the NERT model are described below. Following this discussion, descriptions are provided of several other groundwater flow models that have been developed within the NERT model area and surrounding region.

2.5.1 NERT Steady-State Models

The initial version of the NERT groundwater model was a steady-state model developed by Northgate Environmental Management Inc. (Northgate) that was approved on April 4, 2013 by NDEP for use in capture zone evaluation. This model was calibrated to site conditions in 2008/2009, as documented in the Capture Zone Evaluation Report (Northgate 2010). Refinements to the steady-state model were implemented by Ramboll in phases.

The Phase I Model was documented in an attachment to the 2013 Semi-Annual Remedial Performance Report for Perchlorate and Chromium (ENVIRON 2014a). The first phase of model refinements included: 1) an update of the model to reflect current pumping and injection rates of the GWETS, as well as remediation systems of AMPAC and OSSM; and

2) preliminary refinement of the model representation of stream-aquifer interaction near Las Vegas Wash. The model report included a conceptual water budget for the model area. The Phase I Model was used to support the calculation of GWETS performance metrics presented in the 2013 Semi-Annual Report.

The Phase II Model was documented in an attachment to the 2014 Annual Performance Report (ENVIRON 2014b). The second phase of model refinements included: 1) revision of model hydraulic conductivities to incorporate recent aquifer testing results, and 2) further refinements to the representation of stream-aquifer interactions at Las Vegas Wash. The conceptual water balance was also refined in the Phase II Model report to incorporate additional information and updates to the conceptual model.

The Phase III Model was documented in an attachment to the 2015 Annual Performance Report (Ramboll Environ 2015). The third phase of model refinements included: 1) an extended model domain to the north and east, 2) a revised stream boundary to better represent the lateral extent and streamflow of Las Vegas Wash, 3) refined areal recharge rates, 4) the addition to the model of the newly installed TIMET extraction well field and barrier wall, and 5) a refined representation of evapotranspiration from phreatophytes based on 2014 aerial imagery. The conceptual water balance was also refined in the Phase III Model report to incorporate additional information and updates to the conceptual model for site conditions existing in second quarter 2015.

The Phase 4 Model was documented in an attachment to the 2016 Annual Performance Report (Ramboll Environ 2016b). Major refinements to the Phase 4 model include 1) expansion of the model boundaries to correspond with natural geologic boundaries on the east side, 2) incorporation of weirs installed on the Las Vegas Wash, 3) independent evaluation of the conceptual water balance for the model, 4) incorporation of boring information to better characterize the geology near the AMPAC site and Las Vegas Wash, and 5) updates to the remediation systems of NERT, OSSM, TIMET, and AMPAC to reflect 2016 operations.

2.5.2 NERT Phase 5 Transient Flow Model

The Phase 5 Model is a transient flow model that simulates the site conditions for the period from 2000 to 2015. It was documented in a stand-alone report (Ramboll Environ 2016a). One of the main objectives of the Phase 5 Model was to develop a consistent interpretation of stratigraphy within the model area based on the smaller-scale investigations conducted at individual sites by different parties. In order to assist with this effort, a three-dimensional (3D) geological model was developed that integrated data from many different sources into a consistent overall model of the major stratigraphic features that control groundwater flow in the model area. The groundwater model was constructed using the stratigraphy from the 3D geological model. As part of the modeling effort, a detailed conceptual water balance of the model simulation period was developed. This water balance was used as the basis for estimating the inflows and outflows of groundwater in the model.

2.5.3 United States Geological Survey Model

A regional groundwater model of the valley-fill aquifer system of the Las Vegas Valley was developed by the USGS to evaluate possible groundwater management alternatives related to overdraft problems while maximizing use of groundwater resources (USGS 1996). The model incorporates processes such as land subsidence due to groundwater withdrawal, discharges to washes, evapotranspiration, and spring flow. The four-layered model consists of 60 columns and 72 rows with uniform grid size of 3,000 ft by 3,000 ft. The model was developed in two phases. In the first phase, the predevelopment groundwater conditions, representing a period from 1912 through spring 1972, were simulated. The second phase model simulated the period from summer 1972 through Spring 1981, representing development conditions. As a part of the modeling efforts, a conceptual water budget was compiled for the two simulation phases.

2.5.4 University of Nevada at Las Vegas Model

A groundwater model representing the transport of perchlorate from several contaminated sites to the Las Vegas Wash was developed by a team at the University of Nevada at Las Vegas (UNLV) on behalf of the United States Environmental Protection Agency (USEPA). The modeling work included a data survey and compilation that aided the characterization of the contaminated sites in terms of topography, geology, hydrology, hydrogeology, and perchlorate distribution in groundwater (UNLV 2003). The computer model was developed for saturated conditions using the software Visual MODFLOW 2.8 and was calibrated using WinPEST, an automated calibration tool. The model results included an evaluation of the time of travel and potential perchlorate migration pathways from the contaminant sources to the Las Vegas Wash. In addition to the time of travel and concentration distribution, the transport model also evaluated the influence of domestic and industrial wastewater disposal via the infiltration ponds on the development of the plumes.

2.5.5 Las Vegas Wash Model

A groundwater transport model was developed by NDEP to study groundwater/surface water interactions and perchlorate transport along the Las Vegas Wash (McGinley 2003; NDEP 2003). The purpose of the modeling work was to develop a predictive tool to address temporal distributions of perchlorate in the Las Vegas Wash. MODFLOW was used to simulate groundwater flow, with the Las Vegas Wash simulated using the River Package. Only the alluvium aquifer system was simulated in the model.

2.5.6 Athens Road Well Field Model

A solute transport groundwater model was developed by McGinley & Associates (McGinley) to quantify the efficiency of capture at the AWF (McGinley 2007).

2.5.7 Basic Remediation Company Model

A groundwater transport model for the BMI Common Areas, developed by Daniel B. Stephens & Associates on behalf of the Basic Remediation Company (BRC), was documented in BRC (2009). As part of the modeling effort, historical, present, and future conceptual water balances of the study area were developed. A series of predictive solute transport simulations were also conducted for perchlorate, arsenic, hexavalent chromium, and selenium.

2.5.8 AMPAC Model

On behalf of AMPAC, Geosyntec Consultants (Geosyntec) developed a conceptual and numerical model of groundwater flow in the area north of the former PEPCON facility (Geosyntec 2010). A steady-state numerical model was developed to validate the conceptual model against available site data and to develop quantitative estimates of design parameters and operations to remediate the perchlorate plume in groundwater that originates at the PEPCON site. The model was implemented in MODFLOW 2000. In 2013, transient stress periods were added to the model to simulate site conditions for

October 2012-May 2013, the first nine months of AMPAC extraction system operations (Geosyntec 2013). For evaluating the groundwater flow in the northern portion of the plume and the Shallow WBZ, the numerical groundwater model was refined and recalibrated as presented in a technical memorandum (Geosyntec 2016). The numerical model was updated to refine the Shallow WBZ, the interaction with the Wash, and the definition of the drains (Geosyntec 2016). The updated model was used to evaluate the capture zone of the groundwater extraction systems in the Shallow WBZ and the area of the distal, northern portion of the Shallow WBZ perchlorate, and the interaction with the Wash. The capture zone evaluation results were submitted to NDEP in a technical memorandum (Geosyntec 2017).

3. CONCEPTUAL WATER BALANCE

A quarterly conceptual water balance was developed for the model area for the simulation period of 2014-2018. The conceptual water balance provides an independent evaluation of the inflows and outflows of groundwater within the model domain that can be used to guide model refinement.

The components of the conceptual water balance are described in this section. The first component of the water balance described is groundwater discharge to the Wash. Once the groundwater discharge to the Wash is estimated, a simple water balance of the Wash was carried out to evaluate losing and gaining stream reaches. Following this, the report describes the methods and data sources used to estimate other sources of groundwater discharge (pumping, evapotranspiration, storm drains, and boundary outflows), as well as recharge (trenches, unlined ponds, land use recharge, and retention basins) within the model domain. Finally, inflows and outflow across the boundaries of the model domain are described.

3.1 Groundwater Discharge to Las Vegas Wash

Since groundwater discharge to the Wash cannot be measured directly, it is estimated indirectly based on the perchlorate mass loading in the Wash. For the Phase 5 Model, groundwater discharge to the Wash was estimated using a water balance approach based on the streamflow data (Ramboll Environ 2016a). However, because groundwater discharge represents such a small fraction of the flow in the Wash, it is difficult to accurately estimate groundwater discharge using this approach. A refined approach has been used for the Phase 6 Model based on the perchlorate mass balance in different stream reaches. This refined approach is feasible now because there are significantly more perchlorate mass loading data available currently than there were when the Phase 5 Model was developed.

3.1.1 Methodology

There are five USGS stream gage stations on the Wash that fall within the model domain: Duck Creek Confluence (#09419698), Pabco Road (#09419700), Bostic (#09419747), Homestead (#09419749), and Three Kids (#09419753). See Figure 2-2. Three of the gages (Duck Creek Confluence, Bostic, and Homestead) were installed in September 2016 by USGS within the model simulation period. There are two gages, Las Vegas (LV) Wasteway (#09419679) and Duck Creek (#09419696), located just upstream of the western model boundary (Figure 2-2). The Duck Creek gage is located on a tributary that joins the Wash just upstream of the Duck Creek Confluence gage. To the east of the model boundary, the Wash flows in a subsurface pipe beneath Lake Las Vegas and resurfaces just before the Northshore Road gage (#09419800), located 3 miles downstream of the Three Kids gage (Figure 2-2). Daily streamflow rates for these stream gages downloaded from the USGS website are shown on Figures 3-1 through 3- 7.

To estimate the groundwater discharge to the Wash, the portion of the Wash that adjoins the model domain was divided into five reaches bounded by USGS stream gages: Reaches 1 through 5. Just before the Pabco Road gage, treated effluent is discharged into the Wash at outfalls from the City of Henderson (COH), TIMET, Endeavour, and NERT. The reaches and outfall locations are shown on Figure 2-2 and shown schematically on Figure 3-8. Perchlorate mass loading is calculated at each stream gage. Within each reach, the perchlorate mass discharge is calculated as the difference

in mass loading between the downstream and upstream stream gage. The groundwater discharge into each reach is estimated by dividing the increase in mass loading by approximate perchlorate concentration of groundwater discharging in that reach. A similar approach was presented by Dr. Weiquan Dong of NDEP during 2018 Nevada Water Resources Association (NWRA) annual meeting (February 27 – March 1, 2018).

3.1.2 Perchlorate Mass Discharge to the Wash

Perchlorate mass discharge to the Wash has been estimated quarterly for Reaches 1 through 5 (Figure 3-9, Table 1). Quarterly mass loadings were calculated by multiplying the measured perchlorate concentrations in the Wash at nearby transects by the instantaneous streamflow measurements at the time of sampling. As part of the Downgradient Investigation, AECOM also collected perchlorate samples from various locations in the Wash and the mass discharge estimates they reported are included in Table 1 for comparison (AECOM 2019). The estimated values provided by AECOM are similar to the averages determined from the data collected by NERT, where data are available for comparison. Using the average quarterly loading at each location, the perchlorate mass discharge entering the Wash for each reach was calculated by subtracting the loading estimated at the downstream station from the estimate at the upstream station.

As stated previously, the Duck Creek Confluence, Bostic, and Homestead stream gages were installed in September 2016. To estimate the loading at these locations prior to their installation, the total loading between the nearest upstream and downstream gage locations was distributed based upon the average distribution found in 2017. The loading estimates from 2017 were selected because 2017 was the first year in which there was a complete dataset for all gage stations available and there were not any unusual activities such as weir construction or dewatering, which would affect average streamflow. The loading estimates are shown on Table 1 and Figure 3-9.

Within the simulation period there were a few quarters in which either the streamflow or the analytical data were unavailable for this analysis. For example, for the LV Wasteway station the streamflow data were not available between January 2016 through February 2017, while USGS relocated the gage station. Gage information for LV Wasteway became available in March 2017, meaning only one monthly analytical sample was available for the first quarter 2017 (2017Q1). In this case, the average loading from 2017Q2 through 2017Q3 was used to determine the 2017Q1 value. Then, the average of each quarter from 2015 and 2017 was used to determine the respective quarterly averages for 2016 at the LV Wasteway location.

For each of the Bostic and Homestead locations, analytical sampling began at the end of June 2017 and only one analytical sample was available at each location for 2017Q2. Rather than using a single measurement for the quarterly estimate, the average loading from 2017Q3 and 2017Q4 for Bostic and Homestead were used to determine the value for 2017Q2. To estimate the loading during 2017Q1 for the Bostic and Homestead gages, the averages of 2017Q4 and 2018Q4 were used to best reflect the similar seasonal conditions during the winter months. The average values in 2017 were then used to estimate the distribution between reaches for earlier years, as previously described.

At the Three Kids location, unusually high fluxes were observed between the fourth quarter of 2015 through the end of 2016. Perchlorate loading at the Three Kids gage station is expected to be equal to the loading at the Northshore Road station. Hence, for quarterly loading estimates, the large loading values at Three Kids for 2015Q4 through 2016Q4 were replaced with the quarterly averages measured at Northshore Road. The difference between the loading at Northshore Road and Pabco Road was then distributed to the reaches according to the distribution of loading for Reaches 3-5 observed during 2017, as described earlier in this section. For comparison against the total loading estimated for the Wash, the average loading of the raw data from Three Kids has been included in Table 1.

3.1.3 Perchlorate Concentrations in Groundwater Near the Wash

The measured groundwater perchlorate concentrations near each stream reach were estimated using groundwater sampling data near the Wash in the second quarter of each year. In general, monitoring wells were selected using a 300 ft buffer zone to the south of the Wash. For Reach 1, although WMW6.9S is only well located within the 300 ft buffer zone, its location is too close to Reach 2 for its concentrations to be representative of Reach 1. Instead, PMW-8 which is located further away from the Wash was used. For Reach 2, perchlorate concentration data from groundwater sampling of well PC-155A, PC-155B, PC-156A, PC-156B, PC-157A, PC-157B, WMW6.55S, WMW6.15S were used. For Reach 3, data from COH-2B and WMW5.5S were considered. For Reach 4, data associated with groundwater sampling from LNDMW-1 and WMW4.9S were used. For Reach 5, data associated with groundwater sampling from WMW3.5S, which was the only well located within 300 ft buffer zone, was used. The location of these monitoring wells is shown on Figure 2-2.

The ranges of measured groundwater perchlorate concentrations for Reach 1 through Reach 5 were 86 to 120 micrograms per liter (μ g/L), 1,800 to 3,400 μ g/L, 2,350 to 9,410 $\mu q/L$, 1,700 to 2,000 $\mu q/L$, and 790 to 1,700 $\mu q/L$, respectively, for the model simulation period. The measured perchlorate concentration in groundwater at well WMW3.5S for 2014 (790 $\mu q/L$) is almost fifty percent lower than the concentrations measured for this well in other years during the simulation period. Hence, for Reach 5, the groundwater concentration measured in groundwater at WMW3.5S well in 2015 (1,500 µg/L) was used for 2014.

3.1.4 Groundwater Discharge to Las Vegas Wash

The annual average of perchlorate mass discharge for each reach and each year was determined by using the quarterly estimates given in Table 1 as described earlier in this section and shown in Table 2a. The groundwater discharge was calculated by dividing average annual perchlorate mass discharge by the average groundwater perchlorate concentrations in each reach (Table 2a). For the period 2014 to 2017, in Reaches 1-2 the groundwater discharge was estimated in the range of 166,000 to 211,000 cubic feet per day (cfd), and in Reaches 3-5 the total groundwater discharge was estimated from 447,000 to 510,000 cfd. The estimated groundwater discharge for the period 2014 to 2017 was fairly consistent.

In 2018, due to dewatering activities during weir construction, the estimated groundwater discharge in Reaches 1-2 was lower as compared to the estimates for the period 2014 to 2017. For 2018, the estimated groundwater discharge in Reaches 1-2 and Reaches 3-5 were 24,500 cfd and 506,000 cfd, respectively (Table 2a).

3.1.5 Evaluation of Gaining/Losing Stream Reaches

The estimated groundwater discharge in individual reaches indicate that the Wash is in general a gaining stream. However, the groundwater discharge in Reach 3 (in between Pabco Road and Bostic gage stations) is consistently much smaller than the other reaches. The gaining/losing stream reaches were also evaluated by comparing the stream stage to the groundwater elevations measured in transducers installed by AECOM in June 2017 (AECOM 2019). For this evaluation, for each USGS gage station, a colocated transducer was identified on the southern side of the Wash (Figure 2-2). The Duck Creek Confluence gage was paired with the WMW6.9S transducer. Similarly, Pabco Road, Bostic, Homestead, and Three Kids gage stations were paired with COH2B1, WMW4.9S, LNDMW1, and WMW3.5S transducers, respectively (Figure 3-10). The location of the wells with transducers is shown on Figure 2-2. Like the streamflow data, the stream stage data were also available in the NERT Project database that automatically downloads the real time values from the USGS website. This evaluation showed that at the Duck Creek Confluence gage the Wash is a slightly losing stream (Figure 3-10). The transducer data at Pabco Road and Bostic gage stations clearly indicated that the Wash in Reach 3 is a losing reach. The transducers installed at Homestead and Three Kids showed the Wash is gaining in Reach 4 and Reach 5 of the model domain (Figure 3-10).

The USGS has presented a similar evaluation of various reaches of the Wash during the 2018 Annual NWRA meeting (February 27 – March 1, 2018).² The USGS also determined that Reach 3 is a losing reach; however, USGS reported that Reach 4 was a losing reach in December 2016, but a gaining reach in August 2017.

3.2 Water Balance in Las Vegas Wash

The groundwater discharge estimates described in the previous section were used to develop a simple water balance for the Wash. The data compiled for this evaluation consist of streamflow data from USGS gauging stations, treated wastewater outflows reported by COH, and the reported effluent discharge rates from the treated water from the NERT, TIMET, and Endeavour sites. A schematic of inflows and outflows is shown on Figure 2-2 and illustrated on Figure 3-8. These flow data are summarized in Table 2b. Reach 1 is roughly 60% within the model domain but all other reaches fall completely within the model boundary (Figure 2-2).

The monthly wastewater effluent flow to Las Vegas Wash reported by COH to NDEP has been compiled from 2014 to 2018 for the Kurt R. Segler Water Reclamation Facility.. 3 Annual averages for the COH outfall were approximately 20 cubic feet per second (cfs) in 2014, 21 cfs in both 2015 and 2016, 22 cfs in 2017, and 24 cfs in 2018 (Table 2b). The monthly TIMET outfall discharge rate from 2014 to 2018 was also compiled using data reported to NDEP and then aggregated as annual averages. TIMET outfall was approximately 6 cfs for 2014-16 and 5 cfs for 2017-18 (Table 2b).

For NERT, groundwater is treated using both fluidized bed reactors (FBRs) and an ionexchange treatment system (IX). The effluent from the FBR and IX systems are combined, and they are collectively discharged to the NERT outfall, which is located on a side channel of the Wash near Pabco Road (Figure 2-2). NERT outfall effluent data are

ł

² Presented by Jon Wilson, USGS Hydrogeologist.

³ https://netdmr.ndep.nv.gov/netdmr/public/home.htm

collected daily and maintained by Ramboll within the NERT Project Database. Daily values have been aggregated as annual averages for each year within the simulation period (see Table 2b).

The daily measured streamflow data are very noisy and show large variations over time. To smooth out the streamflow data, 1-year rolling averages were calculated for each of the seven gage stations along the Wash, as shown on Figures 3-1 through 3-7. For LV Wasteway gage station, there was a data gap in 2016 due to the relocation of the gage station by USGS.⁴ Figures 3-1 through 3-7 show that for the period 2014-2018, there were no significant long-term trends in streamflows.

The annual average of measured streamflows for each gage station as determined from the 1-year rolling average is shown in Table 2b. The measured streamflow for each station was calculated as the annual average of flow data below 500 cfs, with flow data above 500 cfs considered outliers. Combined flows from LV Wasteway and Duck Creek gage stations represent the surface water inflow to the model domain (Table 2b). Because no data were available for LV Wasteway in 2016, the surface water flow was assumed to be the same as 2015.

Losses due to evaporation directly from the Wash water surface were accounted for in the water balance using a constant rate. In the Phase 5 Model, an evaporation rate of 78 inches per year was used for the open water bodies. Based on the NDEP comments on the Phase 5 Model, the evaporation rate was revised in the Phase 6 Model by subtracting from the evaporation rate the average rainfall rate of 5 inches per year reported by PRISM (2013). The revised evaporation rate is 73 inches per year, giving total evaporation losses from the Wash from Reaches 1 through 5 of 0.37 cfs, 0.24 cfs, 0.10 cfs, 0.23 cfs, and 0.14 cfs, respectively (Table 2b).

The conceptual net surface water flows at each station were calculated as the sum of surface water flows from upstream, industrial water flows within each reach, if any, and groundwater inflow rates within each reach (as given in Table 2a), minus evaporation from the Wash within each reach. The difference between the conceptual and the measured net surface water flows was equal to or less than 10%, as shown in Table 2b.

3.3 Other Sources of Groundwater Inflow and Outflow

In addition to groundwater discharge to the Wash, other sources of groundwater outflow include groundwater extraction and evapotranspiration. Sources of groundwater inflow include groundwater injection, and both areal and focused recharge.

3.3.1 Groundwater Extraction

Groundwater pumping from 2014 to 2018 was compiled at extraction systems associated with four facilities within the model area: NERT, OSSM, TIMET, and Endeavour (Figure 2-3). Total groundwater pumping rates for individual wells were aggregated from available data for each quarter from 2014 to 2018 (Table 2c).

NERT operates three well fields, including the IWF, the AWF, and the SWF. Endeavour operates three well fields, including the Athens Road Extraction Wells (AREWs), Athens Pen Extraction Wells (APEWs), and Auto Mall Extraction Wells (AMEWs). The AMEW wells were installed in January 2012 to extract a higher volume of perchlorate contaminated groundwater from near the former Endeavour site source area. The AMEWs extract

ł ⁴ Based on email communication with Megan Poff of USGS on 6/28/2018.

groundwater from the Middle WBZ at the depth of 180 ft to 230 ft bgs. All other extraction wells are located within the Shallow WBZ at depth up to 90 ft bgs.

The average pumping rates for each quarter for the IWF, AWF, and SWF ranged from 55 to 74 gallons per minute (gpm) (10,546 to 14,281 cfd), 251 to 471 gpm (48,336 to 90,682 cfd), and 481 to 782 gpm (92,795 to 150,443 cfd), respectively. In addition to NERT's three well fields, AP Area wells (shown on Figure 2-3) started operating in the fourth quarter of 2016, with average pumping rates ranging from 5 to 12 gpm (960 to 2,254 cfd) in each quarter from 2017 to 2018. The pumping data for individual NERT extraction wells obtained from GWETS field sheets were summarized quarterly in Table 2c.

For the OSSM system, the combined quarterly average pumping rates ranged from 127 to 163 gallons per minute (gpm (24,448 to 31,313 cfd) for the simulation period (de maximis 2014, 2015a,b, 2016a,b, 2017a,b, 2018a,b, 2019). For the TIMET system, the range of combined quarterly pumping rates was 22 to 47 gpm (4,302 to 8,952 cfd) (GEI 2019). For the Endeavour system, the range of quarterly pumping rate was 561 to 744 gpm (107,929 to 143,207 cfd) within the model simulation period (Endeavour 2019a). The quarterly pumping rates for individual extraction wells operated by other parties are given in Table 2c.

The total quarterly pumping for individual well fields is also summarized in Table 2d. Figure 3-11 shows that the combined groundwater extraction from NERT, OSSM, TIMET, and Endeavour well fields is fairly consistent between 2014 to 2016. However, the combined pumping rates increased in 2017 as a result of increased pumping at the AWF and SWF.

3.3.2 Groundwater Dewatering (Weir Construction)

During the construction of the Sunrise Mountain and Historic Lateral weirs, construction dewatering from trenches occurred during the first quarter 2018 through the third quarter 2018 (Table 2c). Dewatering at Historic Lateral Weir was completed in early June of 2018, with average pumping rates for the first two quarters of 1,132 gpm (217,906 cfs) and 1,026 gpm (197,506 cfs). Dewatering at Sunrise Mountain Weir was completed in late September of 2018, with average pumping rates for the first three quarters of 2018 of 1,561 gpm (300,601 cfs), 2,091 gpm (402,490 cfs), and 904 gpm (174,088 cfs) (Tetra Tech 2018a-d, f, i, k, m-p, 2019a). Since these trenches were located immediately adjacent to the Wash, a significant portion of the water extracted would have come from surface water leakage. For the water balance, we have assumed that 50% of the water extracted from the dewatering trenches originated from groundwater. The resulting quarterly groundwater extraction rates associated with Sunrise Mountain and Historical Lateral Weir construction dewatering are shown in Table 2d.

3.3.3 Other Groundwater Extraction

A comprehensive review of the Nevada Department of Water Resources (NDWR) groundwater pumpage inventories from 2000 through 2018 was conducted in order to identify any other groundwater extraction occurring within the model area. This was done in response to an NDEP comment on the Phase 5 Model report (NDEP 2017). For each year (2000-2018), any non-environmental extraction wells were identified in the

groundwater pumpage data available from the Las Vegas Valley Water Use Report from Clark County, Nevada, on the NDWR website.⁵

Ramboll identified two non-environmental (one industrial and one commercial) groundwater wells within the model boundary as shown on Figure 3-12, but neither well was actively pumping during the model simulation period (2014-2018). Also shown on Figure 3-12, there are four wells located just outside the model boundary which had minor amounts of pumping during the model simulation period. None of the wells identified on Figure 3-12 were added to the model since none of the wells were located within the model domain and were pumping during the simulation period of 2014-2018.

3.3.4 Evapotranspiration from Groundwater

Like the Phase 5 Model, a reference ET rate of 5.11 feet per year (0.014 feet per day [ft/d]) was applied to the area covered by phreatophytes. According to a USGS study in southern Nevada (USGS 2008), tamarisk (salt cedar) transpire most of their water from April through October. Hence to account for seasonality, the annual ET rate is distributed among four quarters based on the quarterly averages of evaporation rates available for Lake Mead..⁶ The estimated ET rates applied in the model are 0.002 ft/d (Q1), 0.023 ft/d (Q2), 0.022 ft/d (Q3), and 0.010 ft/d (Q4).

The area of vegetation coverage has varied over time due to construction activities near the Wash. The extent of phreatophytes was digitized using a geographical information system (GIS) from aerial photographs provided by SNWA for each year from 2014- 2018.⁷ Figures showing the extent of phreatophytes for each year are provided in Appendix B.

Applying the quarterly ET rate to the extent of phreatophytes resulted in ET that ranged from 22,000 cfd to 365,000 cfd for the model simulation period of 2014-2018 (Table 2d). In addition to the ET in areas covered by phreatophytes, groundwater evaporates from standing water in the gravel pit located west of the APEW wells (Figure 3-13). According to historical topographic maps and aerial photos, standing water was present in the gravel pit during the entire model simulation period. Using an evaporation rate of 73 inches per year, an additional 12,000 cfd of groundwater is lost through evaporation from the gravel pit.

3.3.5 Areal Recharge

Potential sources of areal recharge identified within the model domain for the period 2014-2018 include rainfall infiltration, as well as recharge associated with golf course irrigation (shown in Appendix C, Figures C1-C5).

Rainfall infiltration was evaluated as a potential source of groundwater inflow in the conceptual water balance based on land use.

ł

⁵ [http://water.nv.gov/PumpageInventoryFiles.aspx.](http://water.nv.gov/PumpageInventoryFiles.aspx) The model extent lies within Township 21 South Range 62 East, Township 21 South Range 63 East, Township 22 South Range 62 East, and Township 21 South Range 63 East.

⁶ https://pubs.usgs.gov/sir/2006/5252/table8.html

⁷ Received from Judith Brandt of SNWA.

Undeveloped Areas

The average precipitation rate for the period from 1980 to 2014 near the Site is approximately 5 inches per year.⁸ Historical recharge estimation studies for Nevada groundwater basins have suggested that precipitation recharge is negligible in basins experiencing less than 8 inches of precipitation per year (Maxey and Eakin 1949). The USGS regional study for the Las Vegas Valley (USGS 1996) also reported that precipitation recharge is negligible in the valley, consistent with the findings reported by Maxey and Eakin (1949). Based on an empirical relationship between evapotranspiration and rainfall, given the site climate and under bare land cover the precipitation recharge was estimated as negligible (Sanford and Selnick 2012). Hence, consistent with the previous model versions, precipitation recharge in undeveloped areas was considered negligible in both the conceptual water balance and the Phase 6 Model (Table 2d).

Due to an increase in the area of residential land, the total area of undeveloped land decreased over the period between 2014 and 2018.

Residential Areas

Groundwater recharge in residential areas was assumed to originate from municipal water supply lost to groundwater through irrigation and leaky distribution pipelines. Like the previous model versions, a recharge rate of 9.2 \times 10⁻⁵ ft/d was applied in the residential area within the model boundary.

The area of residential land use within the model domain was estimated from aerial imagery provided by SNWA for each year. During the simulation period, the residential area increased by 12%, from 3.49 X 10⁸ square feet (sq. ft) in 2014 to 3.90 X 10⁸ sq. ft in 2018. The majority of the development occurred between 2016 and 2017, accounting for 6% of the increase in residential area. The expanded residential area results in an increase in the total volumetric recharge rate from approximately 32,000 cfd in 2014 to nearly 36,000 cfd in 2018 (Table 2d).

Industrial Areas

The groundwater recharge rate for industrial areas was estimated using a similar approach as described in the Phase 5 Model report (Ramboll Environ 2016a). The groundwater recharge rate from industrial areas within the model domain is approximately 7.3 X 10-4 feet per day.

The estimated recharge rate was multiplied by the total area of industrial land use for each year, which was constant between 2014 and 2017, at 4.01 X 10⁷ square feet. There was the development of an industrial storage/parking lot area between 2017 and 2018, which accounted for an increase of less than one percent in industrial land use, bringing the total area to 4.02×10^7 square feet in 2018. The resulting volumetric recharge rate for the simulation period was approximately 29,500 cfd, with an increase of less than 40 cfd between 2017 and 2018 (Table 2d).

Chimera Golf Course

The Chimera Golf Club course in the Tuscany Village development has an irrigated area of 128 acres within the model domain (Appendix C, Figures C1 through C5). This course, opened in 2003, was formerly known as the Tuscany Golf Course. Typically, excess irrigation water is applied to turf grass to prevent salt build-up in the root zone. The

ł ⁸ Based on climate data produced by Oregon State University's PRISM Climate Group (PRISM 2013).

amount of turf irrigation water that recharges groundwater can generally be estimated as 25 percent of the annual consumptive use (BRC 2009). The Clark County Area Wide Water Quality Management Plan for 2009 reported the total water usage by the golf course to be 674 acre-feet per year (CCDAQEM 2009). Based on these values, an average recharge rate of 1.78 X 10⁻³ feet per day has been estimated for the golf course area. This corresponds to a total recharge flow of approximately 10,000 cfd (Table 2d). Similar to the Phase 5 Model, it was assumed that the recharge at the golf course was constant for the period 2014 to 2018.

3.3.6 Focused Recharge

Focused recharge from several surface water features in the model domain were evaluated separately and incorporated into the water balance.

COH Wastewater Effluent Ponds

The Kurt R. Segler Water Reclamation Facility (WRF) is presently COH's primary wastewater treatment plant.⁹ Currently, effluent from the WRF is disposed via reuse, discharge into the Wash (as described in Section 3.2), and discharge to the Henderson Bird Viewing Preserve (BVP), as shown on Figure 3-13..¹⁰

Monthly measured flow rates to the ponds of the BVP were obtained from discharge monitoring reports provided by COH.¹¹ To estimate groundwater recharge, it was assumed that the water surface area of the ponds is relatively constant. Daily and monthly evaporation data were available for Lake Mead from 2011 through 2014; however, the average total evaporation using this dataset was 70 inches per year. When compared with the estimate of 73 inches per year for the stream (see Section 3.2), we would expect a higher evaporation rate for the ponds, as a result of having a shallower depth and walls within the ponds that can facilitate additional evaporation. For this reason, we have instead used older historical averages of evaporation that span from 1953 to 1995, published by USGS.¹², with an average evaporation rate of 76 inches per year.

After removing the monthly average of evaporation from the monthly pond outfall estimates, quarterly precipitation estimates were added to the value, resulting in quarterly estimates of recharge at the ponds. Daily precipitation rates reported by USGS were aggregated into quarterly average rates, with the USGS website data automatically being downloaded to the NERT Project database as real time values. The recharge rate from the ponds to the shallow groundwater aquifer for the period 2014-2018 has been determined as quarterly volumetric averages, as shown in Table 2d.

The maximum estimated BVP pond recharge is about 213,000 cfd during 2018Q3 and the minimum recharge is during 2014Q4, with about 53,000 cfd. Generally, the BVP recharge has increased with each year in the simulation period, with an average of 78,000 cfd in 2014, 85,000 cfd in 2015, 121,000 cfd in 2016, 132,000 cfd in 2017, and

ł

⁹ http://www.cityofhenderson.com/utility_services/treatment_facilities.php viewed 11/7/2014

¹⁰ https://www.cityofnorthlasvegas.com/departments/AdministrativeServices/PDFs/BidAds/Bid_1386_4-12- 10_Permit_NV0023647.pdf, viewed 11/7/2014

¹¹ Per data received via email from Howard Analla of the COH, dated 7/09/2013. Data prior to 2008 were obtained via discharge monitoring reports from the NDEP Bureau of Water Pollution Control.

¹² <https://pubs.usgs.gov/sir/2006/5252/table8.html>

184,000 cfd in 2018. The quarterly estimates have been applied for each stress period in the model simulation and are listed in Table 2d.

COH Pond 13

Periodically, COH discharges off-specification effluent to Pond 13 during periods of treatment plant upset. The location of pond 13 is shown on Figure 3-13. According to $COH₁$ ¹³ Pond 13 usually holds up to a week of off-specification effluent, but is generally never used for more than 4 or 5 days at a time. Where available, daily flow data was compiled from 2014 through 2018 and data was then aggregated as quarterly averages. There was a data gap starting in July 2017, with only sparse daily data available through the end of 2018. Where only sparse data was available, we have used the respective quarterly averages spanning all years, 2014 to 2018, in place of the raw quarterly data.

The annual average flow was found to be approximately 8,000 cfd in 2014, 15,000 cfd in 2015, 30,000 cfd in 2016, 16,000 cfd in 2017, and 14,000 cfd in 2018. The minimum flow of approximately 560 cfd occurred in Q1 of 2015 and the maximum flow of over 86,300 cfd occurred in 2016Q4. To calculate the daily infiltration rate, we used a daily water balance and assumed a maximum infiltration rate of 0.4 ft/d, equal to the vertical hydraulic conductivity for alluvium used in the model. Daily precipitation estimates from USGS, available in the NERT Project database that automatically downloads the real time values from the USGS website, were added to the daily flow estimates, and then the corresponding monthly estimate of evaporation (in ft/d) from Lake Mead was subtracted from the total flow, as done with the COH Wastewater Effluent Ponds, and described in the preceding section. Finally, the lesser value of the assumed infiltration maximum, or the available daily water balance was selected as the daily infiltration rate, in ft/d. Quarterly averages were then aggregated and applied as recharge estimates for Pond 13, as shown in Table 2d.

OSSM and TIMET Recharge Trenches

The OSSM and TIMET sites have operated infiltration trenches just downgradient of their extraction wells for disposal of treated groundwater during the model simulation period. For the OSSM and TIMET sites, focused recharge rates at infiltration trenches were estimated as the quarterly average system extraction rates as described in Section 3.3.1 and as shown in Table 2d. The location of OSSM and TIMET recharge trenches are shown on figures presented in Appendix C.

The TIMET area has been estimated as 8.0 \times 10⁴ sq. ft, resulting in an average volumetric recharge rate of approximately 6,800 cfd across the simulation period, with the largest annual average of 8,500 cfd occurring in 2017 and the lowest annual average of 5,100 cfd occurring in 2014. The average recharge rate for TIMET infiltration trenches was 8.50 X 10⁻² ft/d, with a maximum of 1.12 X 10⁻¹ ft/d and a minimum of 5.38 X 10⁻² ft/d.

The OSSM area has been estimated as 1.6×10^5 sq. ft, resulting in an average volumetric recharge rate of approximately 27,500 cfd across the simulation period, with a maximum annual average of 29,000 cfd occurring in 2018 and the minimum annual average of less than 26,000 cfd occurring in 2016. The average recharge rate for OSSM infiltration trenches was 1.70 X 10⁻¹ ft/d, with a maximum of 1.96 X 10⁻¹ ft/d and a minimum of 1.53 X 10 $^{-1}$ ft/d. The quarterly recharge estimates for both the TIMET and

ł

¹³ Phone conversation with Howard Analla of COH on 9/11/2019.

OSSM infiltration trenches have been applied for each stress period in the model simulation and are listed in Table 2d.

NERT Site Storm Water Retention Basins

A recharge flow of approximately 1,800 cfd was assumed for the NERT site storm water retention basins, which has not changed from the Phase 5 Model. In the Phase 5 Model, recharge from storm water retention basins is estimated to be 5.8 X 10-4 ft/d (assuming 50 percent of 5 inches per year rainfall recharges groundwater). This value plus the industrial recharge rate of 7.3 X 10 $⁻⁴$ ft/d was applied to the retention basin areas (as the</sup> retention basins are located in industrial areas). The estimated area for these basins is 1.37 X 10⁶ sq. ft, giving an estimated recharge rate of 1.34 X 10⁻³ ft/d, applied as a constant value across the simulation period.

The total estimated annual groundwater recharge for the simulated period is given in Table 2d. The location of these retention basins is shown on figures presented in Appendix C.

Pioneer Detention Basin

The recharge rate of 4.47 X 10 3 ft/d and volumetric flow of approximately 1,800 cfd from surface water bodies are unchanged from the Phase 5 Model. The approximate area for this basin is 4.0×10^5 sq. ft and the basin is shown on figures presented in Appendix C. The recharge rate was applied as a constant value across the simulation period and the total estimated annual groundwater recharge is given in Table 2d.

Other historical sources of focused recharge, including the BMI Ponds and the TIMET Ponds, were not active during the Phase 6 Model simulation period and are not included in the water balance.

3.4 Storm Drain Inflows and Outflows

The Athens Drainage Channel (ADC), Eastgate Road Storm Drain (EGSD), and subdrains at the Weston Hills development and Chimera Golf Course have been identified as potential sources of recharge/discharge within the model domain.

Storm Conveyance Channels

According to investigations conducted by AMPAC (2007), groundwater containing perchlorate may seep into the ADC and EGSD (Figure 3-14). The ADC was installed in 1996 and extends to a depth of approximately 18 ft below the grade of Athens Road. In order to relieve groundwater pressure from shallow water tables, the ADC was designed to include regularly spaced weep holes comprised of plastic pipes passing through the concrete channel walls into the underlying soil formations beyond the channel. These weep holes, along with certain concrete joints and a limited number of failure points in the concrete structure, allow groundwater to enter the channel and flow to the north as surface water (AMPAC 2007). According to the 2015 Semi-Annual Monitoring and Performance Report for the AMPAC facility (now operated by Endeavour, LLC), groundwater infiltration into the ADC is currently limited, as the shallow water-bearing zone extraction system at Galleria Drive lowers the water table in the vicinity of the ADC (Endeavour 2016a). Endeavour took over operation of the groundwater treatment system from AMPAC in the second half of 2015.

Other sources of flow into the ADC include the EGSD, a subsurface storm drain 12 ft wide and 6 ft high that runs primarily north-south along Eastgate road, and the F6

French Drain, which de-waters the residential area west of Wiesner Way (Endeavour 2016a). In the 2007 AMPAC Performance Report, it was reported that groundwater was entering the EGSD near Eastgate Road between Cape Horn Drive and Sunset Road. This groundwater would have been carried north to the end of the EGSD before discharging to the ADC at the intersection of Boulder Highway and Galleria Drive (AMPAC 2007). Given that groundwater infiltration into the ADC itself is limited due to operation of the AREW extraction wells, groundwater from the EGSD is believed to be the primary source of perchlorate loading to the ADC (Endeavour 2016a). Flow exiting the ADC re-infiltrates back into the ground near the Wash, just upgradient of the former Endeavour reinjection system (AMPAC 2007).

Endeavour currently measures perchlorate concentrations and flow rates in the EGSD and ADC monthly at the approximate locations shown on Figure 3-14. The sampling location in the ADC is downstream of where groundwater seeps into the ADC and upstream of where surface water re-infiltrates to the groundwater system (just east of Wiesner Way) and is therefore representative of flow received from the EGSD, the F6 French Drain, and any potential groundwater infiltration. The perchlorate mass flux upstream and downstream of the junction with the F6 French Drain is approximately equal, suggesting that the F6 French Drain contributes groundwater but little perchlorate to the flow to the Wash. The total groundwater discharge into the channel should be equal to the seepage into the ground at the exit point, except during the storm season when the flow in the ADC would be higher due to stormflows.

As was done for previous versions of the model, monthly flow data recorded at the EGSD and ADC-Main were tabulated from Endeavour's (formerly AMPAC) semi-annual reports and then aggregated as quarterly averages. For the model water balance, the average flow rate across the simulation period at the ADC is 157 gpm, or 30,227 cfd, with approximately 56% of the average ADC flow originating from the ESGD. The minimum cumulative flow at the ADC was 88 gpm (16,990 cfd) in 2014Q4 and the maximum cumulative flow was 207 gpm (39,870 cfd), occurring in 2017Q2. The annual average flows at the ADC were 116 gpm (22,280 cfd) in 2014, 137 gpm (26,310 cfd) in 2015, 158 gpm (30,450 cfd) in 2016, 198 gpm (38,170 cfd) in 2017, and 176 gpm (33,920 cfd) in 2018 (AMPAC 2014, 2015a,b; Endeavour 2016a,b, 2017a,b, 2018a,b, 2019a). The average flow rate at the EGSD monitoring location for the simulation period was 88 gpm, or 16,920 cfd, with a minimum flow of 64 gpm (12,380 cfd) in 2015Q2 and a maximum flow of 117 gpm (22,450 cfd) in 2014Q1. The flows measured at the EGSD contribute a majority of the total flow observed at ADC for all-but-one of the years during the simulation period, accounting for 73% in 2014, 53% in 2015, 52% in 2016, 48% in 2017, and 60% in 2018. For the EGSD, there was an unusually high flowrate measured during 2014Q3 of nearly 230 gpm and this high value was excluded from the averaging process. The quarterly average flows are listed individually for the ADC and EGSD in Table 2d.

The flow of the ADC channel was applied as focused recharge near the Wash as shown on Figures C1 to C5 (Appendix C). The total volume of water recharging groundwater from the channel near the Wash was assumed to be equal to combined flow rates at the EGSD and ADC channel. However, for future mitigation, Endeavour is planning to pump this captured water to Endeavour's treatment plant for perchlorate removal (Endeavour 2019b). To achieve this, infrastructure improvements at the remediation plant will be made that include design and construction of an overhead pipeline to eliminate potential
bottlenecks and installation of a booster pump system on the effluent discharge line (Endeavour 2019b). The construction will reportedly begin in 2020.

Weston Hills and Tuscany Subdrain Systems

The Tuscany development, which surround the Chimera Golf Course, and the Weston Hills development are both equipped with sub-drain systems to help prevent groundwater mounding below buildings. The planned sub-drain locations are shown on Figure 3-15. Ramboll has attempted to obtain the as-built drawings, but has not been successful.

For the Tuscany subdrains, the collection trenches (sub-drains) consist of drain rock and perforated PVC pipe with 6" diameter and 0.02" slot width, oriented across the groundwater gradient (Converse 1998a, 1998b). As part of RI Phase 3 investigations under Modification #2, dated April 19, 2018, the Trust began measuring the flow rate and mass loading in November 2018. Based on data through July 2019, an average discharge of about 50 gpm (9,600 cfd) is expected from the Tuscany sub-drains, which has been applied as a recharge in the area shown on Figures C1-C5 (Appendix C) and presented in Table 2d.

For the Weston Hills subdrains, only sub-drain location information is available, as shown on Figure 3-15. A surface discharge location for the Weston Hills sub-drains has not been identified. It is possible that the discharge from these subdrains is included in the Tuscany sub-drain discharge location.

3.5 CAMU Landfill

The CAMU (Corrective Action Management Unit) landfill is located approximately 10,000 ft west-northwest of the intersection of Lake Mead Drive and Boulder Highway (Figure 2- 3). The landfill was built in 2007 and was closed and capped in 2014. The CAMU landfill is situated adjacent to the BMI Landfill which was operated from 1942 until 1980, at which time it was closed and capped (BRC 2007). Hence, the BMI landfill and the CAMU landfill areas were assigned no areal recharge in the model.

3.6 Boundary Inflows and Outflows

In the Phase 5 Model, inflows and outflows of groundwater enter and exit the model domain from several lateral boundaries and were estimated based on precipitation and elevation data using the approach described in Donovan and Katzer (2000) and shown on Figure 3-16. Based on this approach, the mean estimated recharge rates are 470,000 cfd, 56,000 cfd, and 1,600 cfd coming from the west, south, and north boundaries, respectively (Table 2e). The combined inflows from three model boundaries is expected to be approximately 530,000 cfd (Table 2e). In addition to recharge from precipitation, USGS (2004a) estimated an approximate inflow of 190,000 cfd across the southern boundary from the adjacent Ivanpah Valley and Jean Lake Basins (Table 2d).

However, based on NDEP comments on the Phase 5 Model(NDEP 2017), an assessment of the uncertainty for these inflows and outflows has been evaluated. In response to this comment, four other methods described in Donovan et al. (2009) were evaluated to estimate boundary flow rates: Eakin, Watson, Avon and Durbin, and Epstein methods. Similar to Donovan and Katzer (2000) method, these four methods also assume that precipitation increases with altitude, which increases the groundwater recharge rate. The Maxey-Eakin method is an early model which was revised by Eakin to fit valley-specific conditions. The Watson method uses multiple and simple linear regressions to analyze

the Maxey-Eakins model. The Avon and Durbin method added more data to the Maxey-Eakins method. Their method upheld that the Maxey-Eakins method was sound and corresponded with other calculations done with modern instrumentation. The Epstein et al. (2010) method works in three parts: analysis of historic recharge methods, analysis of statistical uncertainty, and new estimations of groundwater recharge. Hence for the current evaluation, the boundary flow rates estimated from Donovan and Katzer (2000) methods are compared with the Epstein et al. (2010) method. As part of the uncertainty analysis, the maximum, minimum, and the mean flows were estimated for the two methods for all three boundaries, as presented in Table 2e and further described below.

The Epstein et al. (2010) recharge estimates are generally similar to the Donovan and Katzer (2000) method (Table 2e). The Epstein et al. (2010) method resulted in mean inflow of 7,619 cfd versus 1,645 cfd for Donovan and Katzer (2000) in the northern boundary. Similarly, the Epstein et al. (2010) estimate was larger for the southern boundary (132,814 cfd versus 56,151 cfd). In the western zone the Epstein et al. (2010) method yielded 456,196 cfd, while the Donovan and Katzer (2000) method was 467,890 cfd.

The results indicate that there is considerable uncertainty in the recharge rates and there is a rather large range over which recharge rates could be varied to achieve an acceptable groundwater model calibration. In the northern area, recharge rates can vary between 0 to 16,113 cfd. In the south area, recharge rates have a range of 0 to 634,180 cfd. The western area range is 30,749 to 5,299,981 cfd. The individual boundary inflows rates from the calibrated model are expected to fall within the range given in Table 2e.

With the model domain now extended to the Three Kids gage station, it is assumed that the outflow from the eastern boundary is likely to be small as most of the groundwater outflow is expected to be through the Wash as surface water. An outflow of 25,000 cfd has been assumed from the eastern boundary of the model near the Wash. This value was selected because it produces a balance between the inflows and outflows in the basin water balance for the steady-state stress period of 2014 (Table 2d).

The overall estimated inflows and outflows used for the conceptual water balance are summarized in Table 2d. The water balance for the 2014 steady-state period is also shown on Table 2d, with more details on the steady-state period given in Section 5.5.

4. GEOLOGICAL MODEL DEVELOPMENT

This section describes the three-dimensional (3D) geological model that was developed to support the Phase 6 Model. The primary goals of developing the 3D geological model were to: 1) integrate multiple sources of geologic data into a single conceptual representation of geology in the model area; 2) generate a groundwater model grid aligned with geological contact surfaces with appropriately defined hydraulic conductivity zones; and 3) provide a platform for displaying model outcomes and other visualizations of the model area.

4.1 Major Geologic Units

As shown on Figure 1-1 and described in Section 2, the model area is located within the Las Vegas Valley, a basin bounded by the Las Vegas Range, Sheep Range, and Desert Range to the north; by Frenchman and Sunrise Mountains to the east; by the McCullough Range and River Mountains to the south and southeast; and the Spring Mountains to the west.

The focus of the geological model is representing the unconsolidated materials found in the upper 300 to 700 ft of the model domain. In the Las Vegas Valley, eroded Tertiary and Quaternary sedimentary and volcanic rocks comprise the unconsolidated basin deposits, which can be up to 13,000 ft thick (ENSR 2007). Within the model area, the basin fill is up to approximately 5,000 ft thick, based on a gravity study conducted by USGS (1997). The valley floor consists of fluvial, paludal (swamp), playa, and lacustrine deposits surrounded by more steeply sloping alluvial fan aprons derived from erosion of the surrounding mountains. Generally, the deposits grade finer with increasing distance from their source and with decreasing elevation.

Based on the historical description of stratigraphic units found in the model area, the following geologic formations were included in the geological model:

Alluvium

Quaternary alluvial deposits that generally slope north toward Las Vegas Wash are found throughout the model area. The alluvium consists of a reddish-brown heterogeneous mixture of well-graded sand and gravel with lesser amounts of silt, clay, and caliche. The alluvium is comprised of high permeability paleochannels and wash gravels as described in Section 2.1.

Upper Muddy Creek Formation

The Upper Muddy Creek Formation (UMCf) of Pleistocene age occurs in the Las Vegas Valley as valley-fill deposits underlying the alluvium that are coarse-grained near mountain fronts and become progressively finer-grained toward the center of the valley. In borings from the NERT Site, the contact between the alluvium and the UMCf is typically marked by the appearance of a well-compacted, moderate brown silt-to-sandy silt or stiff clay-to-sandy clay, whereas near the Las Vegas Wash, the contact is marked by gray-green to yellow-green gypsiferous clays and silts. Often, a layer of calichified sediments is observed at the contact. Since the coarse and fine-grained materials comprising the UMCf have different hydraulic effects on groundwater flow, the UMCf has been modelled as two primary sediment types:

- Fine-grained facies (UMCf-fg): Fine-grained sediments of clay and silt
- Coarse-grained facies (UMCf-cg): Coarse-grained sediments of sand, silt, and gravel

Within the model domain, the coarse-grained sequences are associated with depositional events originating from the McCullough Range to the South and Southwest, and from the River Mountains to the Southeast. The model incorporates some interbedding of the coarse and fine-grained facies, guided by the USGS description of thinly interbedded depositional sequences present in the Las Vegas Valley (USGS 1989). The process for delineating the extent of the coarse-grained unit is described in the next section.

4.2 Geologic Unit Modeling

A primary application of the geological model is to help define the layer elevations and zone assignments in the groundwater model. Prior to developing the vertical layering of the groundwater model, the geological contact surfaces within the model domain were simulated in a 3D geological model constructed in Leapfrog Hydro (Leapfrog) software. The lateral extent of the geological model is set to approximately match the model area shown on Figure 2-1, though data that extends beyond the boundary can influence the interpolated contact surfaces. The total depth of the geological model ranges from approximately 350 to 400 ft at the Las Vegas Wash to 600 to 700 ft near the southern and eastern upgradient boundaries.

A number of data sources were applied to delineate the lateral and vertical extents of stratigraphic units in the model. In certain locations, in particular along the model boundaries, measured borehole data were sparse, and the extensions of geological contacts were extrapolated outward from the interior of the model domain or adjusted manually by incorporating control lines in the contact surfaces. Cross sections of the final geological model are presented in Figures 4-1 and 4-2.

4.2.1 Ground Surface

For the Phase 5 Model, LIDAR dataset from 2010 was used to for the ground surface elevation data for the model region obtained from SNWA. 14 The LIDAR data was downsampled to a 50 ft resolution grid for input to the geological model. For the Phase 6 Model, the ground surface elevation was updated using the 2016 LiDAR data obtained from SNWA..¹⁵

4.2.2 Alluvium

In the Phase 5 Model, the contact elevation surface between the alluvium and UMCf was developed by incorporating following data sets in the Leapfrog model:

- Elevation values for the alluvium-UMCf contact from the NERT Project Database..¹⁶
- Logs from geotechnical borings drilled near the Las Vegas Wash weirs.
- Geologic cross-sections developed as part of the ongoing NERT Remedial Investigation (RI). These include on-site cross sections presented in the RI Data Evaluation Technical Memorandum (Ramboll Environ 2016c) and westward extensions of these cross sections under development for the forthcoming RI Report.

ł

¹⁴ Received from SNWA via mail on 08/26/2015.

¹⁵ Received from SNWA via mail on 08/28/2017.

¹⁶ The interpretation of the top of UMCf in the NERT Project Database was used to preliminarily define the alluvium interval for each well location. Some locations with suspect data were updated following a more detailed review of available boring logs.

- Geophysical survey data from the Upper and Lower BMI Pond areas (Geovision 2003a) and Las Vegas Wash (Geovision 2003b).
- Structure-contour map and cross-sections developed for the Continuous Optimization Program.
- Cross-sections developed during the Phase III Drilling and Aquifer Testing investigation of areas located between Boulder Highway and Las Vegas Wash (Kleinfelder 2005).
- The surface geology maps for the Las Vegas Southeast quadrangle (Bingler 1977) and Henderson Quadrangle (Bell and Smith 1980), used to define the extent of the alluvium where it pinches out at mountain fronts along the northeast and southwest of the model domain.
- Lithologic logs downloaded from Nevada Department of Water Resources (DWR) Well Logs Database.¹⁷
- Paleochannel delineations outside of the NERT Site and Off-Site Study Area (Northgate 2010).

For the Phase 6 Model, the contact surface was further updated using the dataset from several recent investigations including:

- Borings logged by Ramboll as part of the Phase 2 RI and Phase 3 RI.
- Borings logged by Tetra Tech as part of several pilot study projects including: the Seep Well Field Area Bioremediation Treatability Study, Galleria Drive Bioremediation Treatability Study, AP Area Down and Upflushing Treatability Study, In Situ Chromium Treatability Study, Unit 4 Source Area In-Situ Bioremediation Treatability Study, and Las Vegas Wash Bioremediation Pilot Study.

The boring logs for the RI Phase 2 off-site borings and RI Phase 3 borings were still draft at the time the geological model was developed. There may be minor changes to these contact elevations in future versions of the model. The updated contact elevation surface between the alluvium and UMCf, shown in Figure 4-3, was defined throughout the model domain.

4.2.3 Upper Muddy Creek Formation

The UMCf underlies the alluvium throughout the model domain and is composed of thicker units of the fine-grained facies interbedded with thinner units of the coarsegrained facies. The geological model integrates deeper contact surfaces defining the boundaries between the coarse and fine-grained facies of the UMCf. To simplify modeling of the UMCf, the fine-grained facies was defined as the primary formation. Volumes defining the coarse-grained areas were then subtracted from the fine-grained formation wherever present.

The lateral and vertical distributions of the UMCf-cg units were interpreted from available borehole data and published cross-sections. Due to the limited number of deep boreholes in the model domain, a smaller dataset was available to delineate the UMCf-cg units, and there was greater reliance on cross-sections and professional judgment.

ł ¹⁷ Well Log Database Query Tool: http://water.nv.gov/data/welllog/ Accessed May 5, 2016.

Particular data sets used to derive the deeper stratigraphy in the geological model included:

- Nevada DWR Well Logs Database.
- Boring logs from deep boreholes drilled upgradient of the BMI Upper Pond area as part of the Deep Background Investigation (GES 2007).
- Boring logs from deep borings drilled on the AMPAC and BMI properties.
- Geologic cross-sections developed as part of the ongoing NERT Remedial Investigation (RI). These include on-site cross sections presented in the RI Data Evaluation Technical Memorandum (Ramboll Environ 2016c) and westward extensions of these cross sections under development for the forthcoming RI Report.
- Draft boring logs from recent borings drilled as part of the ongoing Unit 4 and 5 Building Investigation.

The locations of borings that intersect the UMCf-cg are highlighted in Figures 4-4 and 4- 5. Within the geological model domain, the UMCf-cg sequences were associated with depositional series originating from the McCullough Range to the South and Southwest, and from the River Mountains to the Southeast. Conceptually, these sequences were modeled as three separate lateral depositional zones: a "central" region that includes areas on and upgradient of the NERT site, a "west" region that covers portions of the AMPAC site, and an "east" region that extends along the eastern and southeastern model boundary. The three lateral zones were each further subdivided vertically to represent interfingering with the UMCf-fg and disconnected UMCf-cg intervals. The west region was subdivided into three vertical zones, and the east and central regions were subdivided into two vertical zones each. 18 (see cross-section A-A' on Figure 4-1).

The lateral extents of the UMCf-cg depositional zones are illustrated in Figure 4-6. As shown in cross-section Figures 4-1 and 4-2, the different UMCf zones are in lateral and vertical contact in some areas. In certain locations, the youngest and oldest UMCf-cg intervals are vertically separated by bands of UMCf-fg. For all three lateral zones, the UMCf-cg directly underlies the alluvium along the southern and southeastern model boundary. The UMCf-cg does not extend to the base of the model in the west (AMPAC) region.

Figures 4-4 and 4-5 illustrate the total thickness of UMCf-cg and UMCf-fg in the model domain, and the general location of the three depositional zones. Figure 4-4 illustrates the presence of relatively thick deposits along the southern and southeastern model boundaries that thin to the north. Conversely, Figure 4-5 illustrates thickening of the UMCf-fg moving from the southern model boundary towards the wash. The geological model does not fully extend to the base of the UMCf, thus the thickness values shown in Figures 4-4 and 4-5 only represent the thickness of UMCf-cg and UMCf-fg within the model domain.

The xMCf unit, where present in the model, is assumed to be 10 ft thick.

ł ¹⁸ The central zone UMCf-cg intervals correspond to the UMCf-cg1 and UMCf-cg2.

5. FLOW MODELING APPROACH

The results of the conceptual water balance presented in Section 3 and the geological model presented in Section 4 were used as the basis for the development of Phase 6 Model. The key model components that are revised in this version from the Phase 5 Model are described in this section.

5.1 Model Grid

The eastern boundary of the model was extended to Rainbow Gardens Weir to include Three Kids gage station (Figure 2-2). The Phase 5 Model had a uniform grid size of 100 ft by 100 ft. In the Phase 6 Model, the grid was further refined to 50 ft by 50 ft in the well field areas, the Wash areas, and near the unit buildings at the NERT Site. In general, the vertical layering of the MODFLOW grid match layer elevations to geologic contact surfaces wherever possible, while avoiding abrupt changes in layer thickness and elevation. The minimum layer thickness is 4 ft.

Unlike the seven-layer Phase 5 Model, the Phase 6 Model has ten layers, with the top layer designated as Layer 1 and the bottom layer designated as Layer 10. Model Layer 1 represents the alluvium. Layers 2 through 10 represent the UMCf, with hydraulic zones defined to indicate the presence of UMCf-fg (fine-grained) and UMCf-cg (coarse-grained). The UMCf-fg is present throughout the northern portion of the model domain in Layers 3 to 10, and the UMCf-cg is generally present in the central, eastern, and western depositional areas. The layer thicknesses and lithologies represented by each layer are provided in Table 3.

5.2 Simulation Period

For the Phase 6 Model, the simulation period of 2014-2018 was selected because of the availability of concentration data for transport model calibration. There were no major changes to the remediation system during this period, except for pumping rates in the existing well fields (as described in Section 3.3.1).

The model simulation period was divided into twenty quarterly stress periods. Each stress period was assigned recharge rates, evaporation rates, and pumping/injection rates consistent with the conceptual water balance described in Section 3.

5.3 Hydraulic Properties

The summary of horizontal and vertical hydraulic conductivities estimated from various aquifer tests done on different wells in the model domain is presented in Table 4 and Appendix D. The range of hydraulic conductivity values estimated for each geologic unit was used to update the distributions of hydraulic conductivities for various layers in the model as described below.

5.3.1 Horizontal Hydraulic Conductivity

The modeled horizontal conductivity values for each model layer are shown on Figures 5- 1 through Figure 5-8. The conductivity values are within measured ranges for each lithologic formation (Table 4) and are very similar to the Phase 5 Model, except as noted below.

• The conductivity for alluvium in Layer 1 was increased from 40 ft/d to 45 ft/d in the model (Figure 5-1).

- The conductivity of the Wash alluvium was increased (from 200 ft/d to 300 ft/d, from 485 ft/d to 550 ft/d, and from 650 ft/d to 700 ft/d) as shown on Figure 5-1.
- For Layer 10, the conductivity values were updated for better calibration of artesian wells as shown on Figure 5-8. A conductivity value of 30 ft/d was used near the southern model boundary for the UMCf-cg to represent confined conditions for the artesian wells.

5.3.2 Vertical Hydraulic Conductivity

A summary of measured values of vertical conductivity is provided in Appendix D and Table 4. In the model, the vertical conductivities were revised from the Phase 5 Model to better calibrate vertical gradients and heads in the lower layers. The conductivity values that have changed from the Phase 5 Model are noted below.

- For the alluvium, the vertical conductivity was decreased from 0.6 ft/d to 0.4 ft/d.
- In the bottom layers, for UMCf-cg, the vertical conductivity was changed from 0.12 ft/d to 0.012 ft/d to simulate confined conditions for the artesian wells.

In general, the values of vertical hydraulic conductivities in the model are higher than the geometric mean of the estimated values for various geologic formations (Table 4). However, the modeled values are within the range of the observed values except for the xMCf, where the modeled vertical conductivity (0.6 ft/d) is set to a value higher than the maximum observed value $(5.9 \times 10^{-4} \text{ ft/d})$ in order to improve the calibration.

5.3.3 Storage Properties

Storage properties were assigned based on the geologic unit and were adopted from the Phase 5 Model. The storage properties include the specific yield for unconfined layers and the specific storage for confined layers. Effective porosity, used in particle tracking, was set equal to the specific yield. The storage property values assigned are shown in Table 4. The assigned values are generally consistent with the results of aquifer tests conducted in the model area (Appendix D) and with ranges reported in the literature for similar geologic material types (Todd 1980).

5.3.4 Horizontal Flow Barriers

There are several features within the model area that act as barriers to flow that are simulated using the Horizontal Flow Barrier (HFB) package. These features include the slurry walls at the NERT and TIMET sites, the Frenchman Mountain Fault, and the sheet piles associated with weirs installed along the Wash. The horizontal flow barriers are shown on Figures 5-9 through Figure 5-12 and described below.

NERT Barrier Wall

The slurry wall on the NERT Site is located immediately north of the IWF. This feature is simulated as an HFB boundary. The reported range of conductivities used during construction was 4.7×10^{-8} centimeters per second (cm/sec) to 8.0×10^{-7} cm/sec (Vector 2011). This range was similar to the average hydraulic conductivity measured by permeability testing of the barrier wall at four locations of 8.8×10^{-7} cm/sec, as reported in the Capture Zone Evaluation Report (Northgate 2010). For modeling purposes, the value of 8.8×10^{-7} cm/sec was used to represent the barrier wall's hydraulic conductivity. According to the conceptual site model developed by ENSR, the slurry wall is about 1,600 ft long, 3 ft thick, and approximately 60 ft deep, and was constructed to tie into approximately 30 ft of UMCf (ENSR 2005). The layer thicknesses were adjusted in the

model to more accurately represent the slurry wall configuration. An evaluation of the barrier wall integrity demonstrated that the barrier wall is serving its intended purpose (Ramboll 2019a).

The NERT barrier wall is simulated in the top three layers of the model, extending to a depth of 60 feet (Figures 5-9 through 5-11).

TIMET Barrier Wall

The TIMET slurry wall was completed in early 2014. This slurry wall is represented in the model as an HFB boundary. Using information contained in the construction report (GEI 2014), the slurry wall was represented in the model using a hydraulic conductivity of 1x10⁻⁶ cm/sec, a total length of 2,410 ft, and a depth of approximately 60 ft. The thickness of the wall was assumed to be 3 feet.

The TIMET barrier wall is simulated in the top three layers of the model, extending to a depth of 60 feet (Figures 5-9 through 5-11).

Frenchman Mountain Fault

The model domain includes the Frenchman Mountain Fault, located 0.5 miles southwest of the Three Kids Weir (Figure 2-2). This fault is part of the Las Vegas Valley shear zone, a northwest striking, right-lateral strike-slip fault zone (USGS 2005). The last movement on the Frenchman Mountain Fault has been interpreted to be late Pleistocene to early Holocene, approximately 10,000 years before present (GES 2003). According to the geotechnical investigation report for the Three Kids Weir, the upper 30 ft of the Las Vegas Wash floodplain deposits were not offset by the fault in this area (GES 2003). Hence, this fault is simulated in Layers 2 through 10 only. Ramboll is not aware of specific information regarding the hydraulic properties of the fault. For modeling purposes, this fault has been assumed to slightly impede groundwater flow in the units beneath the alluvium. The width of the fault zone has been assumed to be 10 ft with a conductivity value of 0.065 ft/d (which accounts for 10% of the conductivity of Horse Spring Formation).

The fault is simulated in the model for Layers 2 through 10 (Figures 5-10 through 5-12).

Weir Sheet Piles

The weir sheet piles are simulated as HFB boundaries (Figure 5-9 and Figure 5-10). The input parameters for the HFB package remain unchanged from the Phase 5 Model.

5.4 Boundary Conditions

The model boundary conditions include lateral flows of groundwater across the model boundaries, stream-aquifer interaction at the Wash, evapotranspiration, pumping/ injection from wells, as well as areal and focused recharge. Boundary conditions applied for the Phase 6 Model are described below.

5.4.1 Lateral Boundary Inflows and Outflows

Lateral boundary flows in the model were defined based on the conceptual water balance. These inflows were simulated using specified flux boundary conditions (WEL package). In the Phase 5 Model, a general head boundary (GHB) condition was used to simulate lateral boundary flows. The location of the boundary conditions is shown in Figure 5-9 through Figure 5-12.

At the southern boundary in Layers 2 through 10, the boundary conditions with higher fluxes were applied in places where the UMCf-cg unit is present. The boundary inflows through the UMCf-fg are expected to be minimal due to the lower hydraulic conductivity of the finer-grained unit. Hence, the southern boundary fluxes in Layers 3 through 10 where the UMCf-fg is present are assigned a smaller inflow.

The western, northern, and eastern model boundary conditions are also simulated as specified flux boundaries in Layers 1 and 2 of the models (Figure 5-9 and Figure 5-10).

5.4.2 Stream Boundary

The stream boundary condition represents the stream-aquifer interaction at Las Vegas Wash. As shown in Figure 5-13, the stream network includes the Wash, Duck Creek, the C-1 Channel, and a small tributary stream carrying surface water discharges near Pabco Road. Like the Phase 5 Model, the Stream Flow Routing (SFR) package (USGS 2004b) was used to simulate stream-aquifer interaction, as well as evaporation from surface water. The stream boundary was divided into a total of 67 segments. The stream segments are defined as parallel to the direction of flow of the stream, with each weir defined by a separate segment to allow the specification of reduced streambed conductivity at the concrete weir structures (Figure 5-13). The segments are not meant to represent the actual flows within portions of the Wash, rather, they are used solely to allow the tabulation of total flows at the stream gages. Inputs to the SFR package include the evaporation rate from surface water, streambed conductance, stream stage, and streambed elevation. The applied evaporation rate to the stream cells is 0.0166 ft/d.

The streambed conductance is a function of the area of the stream in each grid cell, the thickness of the streambed, and the hydraulic conductivity of the streambed. The areal extent of the stream in each stream grid cell was estimated based on 2017 aerial imagery. The streambed thickness was set uniformly to 1 ft, and the streambed hydraulic conductivity was manually adjusted until the groundwater inflow to each stream segment approximately matched the net groundwater inflow calculated in Table 2a. In general, lower conductivity values were applied upstream of Pabco Road where the streambed slope is less steep. Near the weirs, a smaller value of streambed conductivity was used to reflect the lower conductivity of concrete and other materials used in weir construction. The distribution of streambed conductivity values in the model is shown in Figure 5-13. Streambed elevations were obtained from weir construction information received from SNWA (Table 5).¹⁹ Stream stage and streambed elevation profiles are implemented in the stream boundary as shown in Figure 5-14. The depth of the Wash is kept fixed during the simulation.

All weirs have sheet piles installed into the alluvial deposits (and in some cases into the UMCf). The weir sheet piles act as a barrier to groundwater flow, and force groundwater to pass beneath or around the sides of the sheet pile. Like the Phase 5 Model, the sheet piles were simulated using the HFB boundary condition. Table 5 shows which model layers include simulated sheet piles for each weir.

5.4.3 Evapotranspiration from Groundwater

The area of phreatophytes located along the Wash was refined in the model by digitizing areas of riparian vegetation visible in aerial images from 2017 received from SNWA. The

ł ¹⁹ Received from SNWA via email on 02/4/2016.

resulting zones of evapotranspiration are shown on Figure B1-B5 in Appendix B. A maximum evapotranspiration rate of 5.11 ft per year was assigned with an extinction depth of 15 ft, as was used in the Phase 5 Model.

As discussed in section 3.3.4 above, in addition to groundwater loss through phreatophytes, there is an additional loss of groundwater through evaporation from standing water in the gravel pit located west of the APEW wells (Figure 3-13). An evaporation rate of 73 inches per year (as described in section 3.2) was assigned in model Layer 1 in the 76 grid cells where standing water is visible within the gravel pit. An extinction depth of 1 ft was used in these cells to simulate surface water evaporation.

5.4.4 Groundwater Pumping and Injection

Groundwater pumping and injection were simulated using either the standard Modflow Well Package (WEL) or the Multi-Node Well Package (MNW1), with average quarterly rates for the period 2014-2018 for each location shown on Figure 2-3. For each well, the pumping rates, well screen intervals, and corresponding model layers are shown in Table 2c. Treated groundwater pumped into the OSSM, and TIMET recharge trenches was applied as areal recharge in the model at the locations shown in Appendix C.

The WEL package was used to simulate extraction wells screened within one layer. In addition, the WEL package was also used for wells in the IWF, the OSSM and the TIMET well fields because using the MNW1 package caused convergence issues. For wells screened in more than one model layer, pumping was assigned to individual layers based on the proportion of screen in each layer.

The MNW1 package was used to simulate SWF, AWF, and Endeavour wells where the well screen falls in more than one model layers. This package simulates the allocation of pumping from different model layers internally based on the screen interval and the hydraulic conductivity of the formation (Konikow et al. 2009). Well losses were not simulated, and the well radius was set to less than the effective radius of the model grid cell.

5.4.5 Areal and Focused Recharge

Areal and focused recharge has been updated as described in the conceptual water balance in Sections 3.3.5 and 3.3.6, and shown in Appendix C (Figures C1 Through C5) and Table 2d.

5.4.6 Drainage

In addition to the EGSD and ADC, the Golf Course subdrains and Weston Hills subdrains were added to the Phase 6 Model as described in the conceptual water balance in Section 3.4. The location of these subdrains is shown on Figure 3-14 and Figure 3-15. These drainage channels are simulated in the model using the drain (DRN) package.

Since the location of the F-6 French drain is unknown, a hypothetical drain boundary was added to the west of the ADC in order to match the flow observed in the ADC, as shown on Figure 5-9. The reference elevations were initially defined to be in the middle of Layer 1 under the water table. Then the drain conductance and reference elevations were adjusted to match the measured flow (Table 2d).

5.5 Model Initial Heads and Transient Simulation

The initial condition for the transient simulation was defined using a steady-state simulation representing average groundwater conditions in 2014. For this purpose, an additional stress period (in steady-state) is added to the model at the beginning of model simulation. For this stress period, the boundary conditions and various inflow parameters are assigned based on annual averages in 2014, as shown in Table 2d. Thus, the model has a total of twenty-one stress periods, with the first stress period being steady-state and the remainder transient.

5.6 Modeling Software

The Phase 5 Model was simulated using the USGS modeling code One-Water Hydrologic Flow Model (MODFLOW-OWHM) version 1.00.00 (Hanson et al. 2014) in order to represent the installation of slurry walls and sheet piles below weirs that occurred during the simulation period. With the change in simulation period, there is no need to include a representation of weir installation in the Phase 6 Model, so the model was developed using the simpler USGS modeling code MODFLOW-NWT Version 1.1.4 (Niswonger et al. 2011). Most of the model input files were generated using the Groundwater Vistas interface (Version 7.24 build 70), with the exception of the SFR input file. A complete set of model files is included in Appendix J.

6. FLOW MODEL CALIBRATION

The Phase 6 Model was calibrated by varying the model parameters so that the simulated model results are consistent with the observed data, the conceptual water balance, and the overall conceptual model of groundwater flow in the model area. This section describes the model calibration process and evaluates the quality of the model calibration.

6.1 Calibration Objectives

The model was calibrated using a combination of automatic calibration and a trial-anderror approach. The model calibration objectives were as follows:

- To match the major flow components of the conceptual water balance (including pumping, surface recharge, and flow in the drainage channel), as given in Table 2d.
- To match simulated groundwater discharge to the Wash within 15% of the conceptual estimates.
- To obtain a volumetric mass balance error of less than 1%.
- To match head targets with a residual standard deviation to range ratio of less than 10% for non-artesian wells and less than 20% for artesian wells (since artesian wells are hard to calibrate with a regional model).
- To obtain predicted capture zones for each of the well fields that are consistent with the conceptual site model.

6.2 Model Calibration Targets

The measured groundwater levels are referred to as model calibration targets. During model calibration, model parameters were adjusted so that the simulated water levels were as close as possible to the head calibration targets. Two types of head calibration targets were used in the model: 1) groundwater elevation targets and 2) head difference targets. The model was also calibrated to the streamflow estimates and the pumping rates as given in Table 2d. The methodology and data sources for these targets are as described below.

Groundwater Elevation Targets

The average quarterly groundwater level data for 2014-2018 were compiled from the NERT Project Database of groundwater elevations. There are several sources including:

- Data collected by the Trust in 2014-2018:
	- − Data from the Groundwater Monitoring Program (Ramboll Environ 2014, 2015, 2016b; Ramboll 2017, 2018c, 2019b);
	- − Data from Phase 1-3 of the RI;
	- − Data from treatability studies;
		- o Zero-Valent Iron (ZVI)-Enhanced In-Situ Groundwater Treatability Study (Ramboll 2018a);
		- o AP Area Treatability Study (Tetra Tech 2018h);
		- o Galleria Drive Bioremediation Treatability Study (Tetra Tech 2018l);
		- o Groundwater Bioremediation Study (Tetra Tech 2016);
- o In Situ Chromium Treatability Study (Tetra Tech 2018e);
- o Las Vegas Wash Bioremediation Pilot Study (Tetra Tech 2017b, 2018r);
- o Seep Well Field Area Bioremediation Treatability Study (Tetra Tech 2019b);
- o Soil Flushing Treatability Study (Tetra Tech 2017a); and
- o Vacuum Enhanced Recovery Treatability Study (Tetra Tech 2018g);
- Data received directly from BMI complex parties in 2014-2018 for use in the Annual Remedial Performance Reports for Chromium and Perchlorate for the NERT Site (Ramboll Environ 2014, Ramboll Environ 2015, Ramboll Environ 2016b; Ramboll 2017, Ramboll 2018c, Ramboll 2019b);
	- − Data provided by Endeavour;
	- − Data provided by OSSM; and
	- − Data provided by TIMET;
- Data received from SNWA in 2014-2018 for use in the Annual Remedial Performance Reports for Chromium and Perchlorate for the NERT Site (Ramboll Environ 2014, Ramboll Environ 2015, Ramboll Environ 2016b; Ramboll 2017, Ramboll 2018c, Ramboll 2019b); and
- Data received from BRC in 2014-2015 for use in the Annual Remedial Performance Reports for Chromium and Perchlorate for the NERT Site (Ramboll Environ 2014, Ramboll Environ 2015, Ramboll Environ 2016b; Ramboll 2017, Ramboll 2018c, Ramboll 2019b).

After data compilation was completed, data were evaluated based on the quality of the data source, the date of measurement, the location of the measurement within the model grid, and the amount of data available in the period between 2014 and 2018. Dry wells were not included. In total, there were 1039 wells and 12,869 groundwater elevation measurements considered for 2014 to 2018. For the first time step (0.5 day), which corresponds to the flow model steady state simulation, the average groundwater elevations in 2014 were used as targets and 465 annual targets were considered. For the transient model, the targets were averaged on a quarterly basis in order to compare them against the simulated quarterly stress periods. There were 6,656 quarterly water level targets used for 2014 - 2018 transient model calibration.

The frequency of measurement varied with time and location. In order to address the quality of available data, weights were applied to targets. A simple weighting scheme was used based on the statistical principle that the accuracy of the mean is proportional to the square root of the number of samples. Weights of each quarter target were set to one-half the square root of the number of quarters with one or more head measurements for that location and quarter. The maximum value of weights is set to be 1 for wells with frequent data.

The model layer in which target wells are screened was determined from well construction information stored in the NERT Project Database.

The representative groundwater elevation targets used for model calibration are tabulated in Appendix E and are shown on Figure 6-1 for 2014 targets and Figure 6-2 for 2018 targets. As a result of additional investigations conducted after 2014, there are additional targets in 2018.

Head Difference Targets

The target locations for the vertical head differences are shown in Figure 6-3. Representative head difference/vertical gradients at various well clusters used for model calibration are tabulated in Appendix F. Tables F-1a to F-1e show vertical gradients that were calculated from manual groundwater elevations from 2014 to 2018, respectively, and Tables F-2a to F-2b show vertical gradients that were calculated from locations with automatic transducers from 2017 to 2018, respectively.

Vertical gradients were calculated from manual groundwater elevations measured as part of routine groundwater monitoring within the RI study area. Locations used for this portion of the analysis included all available and applicable wells in the NERT Project Database, using the same selection method for head targets. First, clusters of wells within a 100-ft radius (and a 50-ft depth radius) were identified. A cluster of wells also needs to have at least 50-ft vertical distance between the highest midscreen elevation and the lowest midscreen elevation. In total there were 84 clusters identified (Figure 6- 3). Quarterly vertical gradients were then calculated for each well cluster from quarterly averaged groundwater elevation data in 2014-2018. For each cluster, the well which has the highest elevation was identified as "Top" and was used as the reference well. Vertical gradients were calculated for other wells in the cluster by using the following equation.

vertical gradient for well $A = \frac{\text{groundwater elevation of reference well} - \text{groundwater elevation of well}}{\text{midscreen elevation of reference well} - \text{midscreen elevation of well}}$

A positive vertical gradient means the water is flowing downward. The midscreen elevation, midscreen depth below ground surface, model layer, WBZ, and lithology in screen interval for each well can be found in Appendix F. In 2014Q1, 2014Q3-4, 2015Q3-4, and 2017Q1, no vertical gradient could be calculated due to lack of data. For each quarter, clusters that have data available are shown in Tables F-1a to F-1e. A cluster is shown when there are head data available for at least one well in the cluster.

Additionally, vertical gradients were calculated from locations with automatic transducers using the same approach. There were 8 clusters identified (Figure 6-3). Data were available from 2017 to 2018 (Tables F-2a and F-2b). Locations used for this portion of the analysis were wells in the Eastside and Downgradient Study Areas, with transducers installed in 2017 in support of the groundwater monitoring program. Data analyzed in this effort included all available well transducer data points from 2014 to 2018 in the All Wells Database on June 7, 2019. Automatic transducer elevation data was averaged across the quarterly time interval, and manual elevations were used as a quality control check. The vertical gradient was calculated for all Alluvium/UMCf well pairs where data was available for both wells.

Other Calibration Targets

Boundary inflows and groundwater discharge rates to the Wash developed in the conceptual water balance were used qualitatively as calibration targets. Additionally, groundwater extraction rates were used as qualitative calibration targets. The estimated boundary flow rates and pumping rates are provided in Table 2d.

6.3 Model Calibration

The Phase 6 transient model was manually calibrated to the water levels and the conceptual water balance for the simulation period (2014-2018) by adjusting the fluxes at the specified flux boundary located at the southern and western boundaries of the model domain. The horizontal hydraulic conductivity values are modified slightly from the Phase 5 Model for better calibration as previously described in Section 5.3.1. The conductivity for alluvium in Layer 1 was increased to 45 ft/d. In the Phase 5 Model, an alluvium conductivity of 40 ft/d was used. This change was made to improve head calibration in shallow water bearing zone wells. The conductivity of the Wash alluvium has been increased (from 200 ft/d to 550 ft/d, from 485 ft/d to 550 ft/d, and from 650 ft/d to 700 ft/d) in the Phase 6 Model as shown on Figure 5-1. This helped improve the match between the conceptual groundwater discharge to Wash and the simulated discharge volume. For Layer 10, the conductivity values were updated for better calibration of artesian wells as shown on Figure 5-10. A conductivity value of 30 ft/d was used near the southern model boundary in UMCf-cg to represent quasi-confined conditions for the artesian wells.

The vertical hydraulic conductivity values were modified throughout the model domain to improve the calibration at the head targets. The vertical conductivity value of the alluvium was decreased from 0.6 ft/d in the Phase 5 Model to 0.4 ft/d. Like the previous model versions, the vertical conductivity in the rest of the geologic units (paleochannels, Las Vegas Wash sediments, UMCf-fg, xMCf, and UMCf-cg) was defined by multiplying the horizontal conductivity by a vertical-to-horizontal anisotropy ratio of 0.1. The calibrated vertical conductivity values in the model are above the estimated geometric mean but are within the range of measured values (Table 4 and Appendix D).

In the model, the hydraulic conductivity assigned to the streambed was adjusted during calibration to obtain a good match between simulated groundwater discharge in Reaches 1 through 5 and the estimated groundwater discharge in the conceptual water balance. The fluxes at the western inflow boundary were adjusted during calibration to match the groundwater flux to the Wash in the stream sub-reaches. The calibrated conductance applied for each stream segment is given in Figure 5-21. The net groundwater outflows from the calibrated model into various reaches of the Wash is within 15% of the values estimated in the conceptual water balance shown in Table 2d.

Convergence criteria of 0.01 ft on head and 500 cfd on flow were specified for the model simulations. The volumetric mass balance error (difference between the total groundwater inflow and outflow simulated by the model) was monitored during model calibration as a check on the model solutions and to identify errors in the model design.

6.4 Model Evaluation

The calibration of the flow model is generally good based on a comparison of simulated and conceptual water budget, simulated and observed heads, and simulated and conceptual groundwater discharge to the Wash. Table 6 presents the major flow components of the simulated water balance of the Phase 6 Model. The overall mass balance error of the final calibrated model was negligible.

There is generally a good match between the conceptual water balance and the simulated water balance. The combined simulated boundary inflows are approximately 840,000 cfd for the simulation period. There is uncertainty in the conceptual measurement of boundary inflows as presented in Table 2e. Thus, the simulated value of boundary flows is reasonable as compared to the conceptual range of combined boundary inflows of 1,100,000 cfd (Table 2d). The total simulated pumping rates for individual extraction wells matches the measured pumping rates given in Table 3. Although there is uncertainty in the conceptual estimates of groundwater discharge to the Wash, the simulated groundwater discharge to the Wash is within 15% of the conceptual estimates (Table 6).

Table 7 provides a summary of target residual statistics for 6,656 observations evaluated for the calibration of the Phase 6 Model. The overall average target residual in the calibrated model is -3.02 ft. A negative residual value indicates that the simulated head is higher than the observed head. There is a root-mean-square (RMS) error of 5.9 ft with the range of observation of 435 ft. The residual standard deviation of the model targets is 5.2 ft. A model is considered well calibrated when the ratio of residual standard deviation and the range of observed groundwater head elevation values is less than 10 percent (Hill and Tiedeman 2007). For the Phase 6 Model, this ratio is about 1.4% percent.

As the deep artesian wells are hard to calibrate within a regional model, for the Deep WBZ (Layers 9 and 10), the calibration statistics of groundwater elevation targets have been evaluated separately (Table 7). The average target heads in these layers is about 10 ft lower than the measured values (target residual of 10.08 ft). The positive value of head residual indicates that the simulated heads are underestimated. There is RMS error of 16.1 ft with the range of observation of 339 ft. The ratio of residual standard deviation and the range of observed heads in the Deep WBZ is 4.8%. The calculation of the target residuals for each target location is shown in Table G-1 (Appendix G).

Figure 6-4 and Figure 6-5 show a scatter plot characterizing the match between modeled and observed groundwater heads and head differences at wells used as calibration targets, respectively. The plots illustrate that there is generally good agreement between modeled and observed heads, with most points falling close to the 1:1 correlation line. The "goodness-of-fit" $R²$ value is 0.99, demonstrating an acceptable fit to the observed heads. The head difference/vertical gradient in Figure 6-5 also demonstrates an acceptable fit except a few outliers. There is in general an upward vertical gradient of heads within the model domain as presented in Appendix F. The vertical gradient results were highlighted in yellow for wells where the simulated vertical gradient had an opposite direction compared to the observed vertical gradient (Appendix F). The target locations for the vertical head difference are shown on Figure 6-3. The well clusters where the simulated vertical gradient had an opposite direction compared to the observed vertical gradient in any stress period of the model simulation are shown on Figure 6-3. For the transducer data, at every location the simulated vertical gradient direction matched the observed vertical gradient (Appendix F-2a and F-2b).

There is a good fit at the vast majority of targets, with differences between observed and simulated head generally less than 10 ft (Table G, Appendix G) for the Shallow WBZ. Table G shows the distribution of targets in between Shallow, Middle, and Deep WBZs. The match between simulated and observed heads in these zones appears reasonable. For the treatability study wells, the model calibration only used the baseline data points. Figure 6-6, Figure 6-7, and Figure 6-8 show the head residuals in Shallow, Middle and

Deep WBZs, respectively. The head residuals shown on these figures are averaged over the model simulation period 2014-2018.

Simulated groundwater discharge to the Wash in the stream reaches also approximately matches the conceptual flow for the water balance period. The observed versus simulated streamflow at various USGS gages is shown on Figures 6-9 through 6-13. The overall simulated streamflow at each gage station is approximately similar to the 1-year rolling averages of the measured streamflow values for the period 2014-2018.

The observed versus simulated groundwater discharge in ADC drains is shown on Figure 6-14. The overall trend of simulated discharge in ADC matches reasonably with the measured discharge data as given in Table 2d. Due to limited information on these subdrains, the conceptual estimate of discharge in the sub-drains is approximate. As a result, the simulated discharge from Weston Hills and Tuscany Golf course sub-drains was not evaluated.

In order to evaluate the capture zones predicted by the model, a particle tracking simulation was performed. Capture zones were estimated using particle tracking for quasi-steady-state conditions existing in 2017. The capture zones for the groundwater extraction systems in Shallow, Middle, and Deep WBZs simulated using the Phase 6 Model are shown on Figures 6-15, 6-16, and 6-17. A qualitative comparison of these capture zones with those based on the Phase 5 Model presented in the 2017 Annual Report indicate that the general configuration of the capture zones for each extraction system remains consistent with the previous model version.

The model is further evaluated by comparing the observed versus simulated heads for Q2-2018 as presented on Figure 6-18. The observed potentiometric contours are from 2018 Annual Report (Ramboll 2018c). The simulated heads are the water table contours exported from the model at the end of 19th stress period representing Q2-2018.

6.5 Sensitivity Analysis of Flow Model

A sensitivity analysis was conducted to evaluate the effect of model input parameter uncertainty on significant model outputs. The model input parameters evaluated included: 1) hydraulic conductivity of major geologic units, 2) surface recharge rates, 3) streambed conductance, 4) storage parameters, and 5) southern and western boundary inflows, and 6) evapotranspiration. The sensitivity analysis was performed by changing each of these model input parameters by a factor of 0.75 and 1.25, running the model, and recording the outputs of interest.

The model outputs evaluated for the sensitivity analysis included: 1) calibration statistics based on target water levels, 2) the difference between simulated and actual discharge to the Wash, and 3) the difference between simulated and actual pumping at extraction wells. The results of the sensitivity analysis for these outputs are shown in Table 8, with the exception of the difference for extraction well pumping, which showed no effective difference in the root mean square error (RMSE) $(3.22 \times 10^5 \text{ cfd})$ as a result of increasing or decreasing any of the input parameters.

Hydraulic Conductivity

The sensitivity of the model calibration to changes in the value of the hydraulic conductivity was evaluated for the alluvium (both outside of and within the paleochannels), the UMCf-fg, and the UMCf-cg. As shown in Table 8, the model calibration is relatively insensitive to vertical conductivities in each of the lithologic units, with most effects occurring within the UMCf-cg. For the horizontal conductivity parameters, the sensitivity analysis results suggest that a lower sum of squared residuals, RMSE, and residual mean for the groundwater elevation could be achieved by slightly increasing or decreasing hydraulic conductivities from the values used in the calibrated model. However, this cannot be done without causing a significant increase to the RMSE for the discharge to the Wash. The only sensitivity run where groundwater elevation statistics were reduced without significantly increasing the RMSE for the discharge to the Wash was the result of increasing the horizontal conductivity within the paleochannels. For this parameter, the associated RMSE for the discharge to the Wash is higher by only 0.4 cfs, which is considered negligible. Further, the calibrated conductivity for the paleochannels is already at the upper end of the range of reasonable values, as presented in Table 4, which indicates that it would not be reasonable to increase the conductivity of the paleochannels to achieve lower statistics.

Surface Recharge

The sensitivity of surface recharge rates was evaluated by uniformly varying the recharge rates shown in Table 2d by factors of 0.75 and 1.25. As shown in Table 8, all three calibration target statistics changed when the recharge rates were modified, indicating that recharge rates are sensitive model parameters. As with the results for horizontal conductivities, although there were lower statistics associated with decreasing the recharge rates by 25%, doing so would result in a much larger RMSE for the discharge to the Wash. Conversely, increasing the recharge rates by 25% results in higher groundwater elevations, but a slightly lower RMSE for the discharge to the Wash.

Streambed Hydraulic Conductivity

Streambed hydraulic conductivity, which controls stream-aquifer interaction, is an uncertain model parameter due to the lack of any direct measurement data. As shown in Table 8, the model calibration statistics based on groundwater elevations or observed pumping are not sensitive to this parameter. As expected, streambed conductivity does influence the rate of groundwater outflow to the Wash as indicated by the RMSE values.

Storage Parameters

The effect on calibration statistics was evaluated by changing specific yield and specific storage values by factors of 0.75 and 1.25. The model results are not especially sensitive to storage parameters, with the exception of the RMSE for the discharge to the Wash, which increased from about 1.9 cfs to about 2.2 cfs with either an increase or decrease to the calibrated storage values (Table 8).

Boundary Inflows

Boundary inflows are uncertain because they cannot be measured directly. As shown in Table 8, the model calibration is sensitive to parameters controlling the southern boundary inflow but not as sensitive to changes in the western boundary inflow. Like other input parameters in the sensitivity analysis of flow, the effect on calibration statistics was evaluated by changing each of these boundary inflows by factors of 0.75 and 1.25. A decrease to the southern and western boundary inflows results in lower statistics for the groundwater elevation, but a significant increase in the RMSE for the discharge to the Wash for both boundaries. Further, when the southern boundary inflow is decreased, the pumping at the shallow OSSM, IWF, and TIMET well fields that can be simulated by the model is reduced from the actual pumping rates, resulting in a difference in total pumping of 4.05×10^6 cfd. Thus, the southern boundary inflow

cannot be reduced while still simulating the observed pumping rates. With higher southern boundary inflows, the overall head residual increases.

Evapotranspiration

Statistics for the groundwater head calibration, discharge to the Wash, and well extraction rates are generally not sensitive to changes in ET. Given an increase in ET rates of 25%, the RM for groundwater heads is slightly less than the calibrated value; however, the RMSE for discharge to the Wash increases significantly. Conversely, a decrease in ET rates of 25% results in higher groundwater heads, but a slightly lower RMSE for discharge to the Wash (Table 8).

7. CONCEPTUAL TRANSPORT MODEL

The model is designed to include all potential source areas of perchlorate and other NERT COPCs, so that the model can be used to better understand the relative contribution of different source areas to discharges into the Las Vegas Wash and impacts to other receptors. The current version of the model focuses on perchlorate transport, but a future version will also include other major NERT COPCs.

A conceptual description of perchlorate fate and transport and a conceptual mass balance are provided in the following sections. The description of perchlorate fate and transport focuses on the NERT perchlorate plume investigated as part of the NERT RI. However, the conceptual perchlorate mass balance includes both the AMPAC and NERT plumes.

7.1 Perchlorate Fate and Transport at the Site

Perchlorate concentrations within the core of the NERT plume in the RI Study Area groundwater have been declining very slowly with intermittent rebounding episodes despite active pump and treat remediation for over three decades (Ramboll 2018c). As discussed in the Phase 1 Remedial Investigation Data Evaluation Technical Memorandum (Ramboll Environ 2016c), perchlorate present in the UMCf is likely migrating upwards into the alluvium as a result of back diffusion and upward flow caused by a natural upward vertical gradient. This is suspected to be the primary reason for the persistently elevated concentrations of perchlorate in several monitoring wells in the downgradient area of the NERT site. The forthcoming RI reports will describe the effects of back diffusion and upward flow in more detail.

Typically, diffusion rates are relatively low compared with groundwater flow rates and, as a result, the effects of diffusion are usually insignificant (as compared to advection) at the scale of site characterization and remediation activities. However, diffusion can be significant when groundwater flow rates are very low (due to low conductivities) and concentration differences exist for long periods of time. Back diffusion has been identified as a significant process controlling the time required to remediate dissolved plumes at complex sites (NRC 2013). Historically, when site discharges containing high concentrations of perchlorate were migrating downgradient in the alluvium, vertical transport from the alluvium to the UMCf would have occurred creating a significant mass of perchlorate in the uppermost reaches of the UMCf. As perchlorate concentrations in the overlying alluvium decreased with the onset of pumping activities and natural flushing, perchlorate mass that accumulated in the UMCf would migrate upward from the UMCf into the alluvium via back diffusion and upward groundwater flow.

These very slow declines of perchlorate concentrations are consistent with the UMCf acting as an on-going source of perchlorate to the alluvium. In some areas, upward vertical hydraulic gradients have been observed and are thought to be causing upward flow of groundwater and perchlorate transport from the UMCf to the alluvium. Since the low vertical hydraulic conductivity of the UMCf-fg limits the vertical flow rates, back diffusion is likely causing significant long-term transport of perchlorate from the UMCf to the alluvium and the effects of back diffusion are likely to be significant within the model area. A more detailed description of conceptual perchlorate fate and transport has been provided in the Phase 5 Model documentation (Ramboll Environ 2016a).

7.2 Fate and Transport Processes

Based on the conceptual model for the site, advection and dispersion are prominent transport mechanisms at the site. As perchlorate is a highly conservative tracer in the soils of the contaminated area and in-situ biodegradation is limited by the absence of an electron donor (i.e., carbon source), adsorption and biodegradation are negligible components in the transport model. Density-driven flow has been evaluated for incorporation into the Phase 6 transport model. As stated above, matrix diffusion is important process in order to simulate recent conditions at the site. In the Phase 6 Model, matrix diffusion is simulated using the dual domain approach. The density-driven evaluation and dual-domain approach are described below.

Density-Driven Flow Evaluation

The density of groundwater is increased in areas with high total dissolved solids (TDS) concentrations, which induces groundwater flow from high density areas to lower density areas. This transport process was likely relevant in areas that historically had high TDS. A separate evaluation of density-driven flow is presented in Appendix A. It is clear from the evaluation that the concentrations have to be greater than 20,000 milligrams per liter (mg/L) for the density effect to be significant. There are limited areas within the model with high enough TDS concentrations for density effects to be significant. In general, density effects do not need to be considered, since density-driven flow does not appear to be the dominant transport process under current conditions. More details are provided in Appendix A.

Dual-Domain Approach

Diffusion is the transport of chemicals due to concentration differences between regions (e.g., the alluvium and UMCf). Accurate simulation of diffusion typically requires a very fine discretization, which is not feasible for regional-scale models. An alternative approach is the dual-domain mass transfer method, which offers a practical solution to modeling perchlorate fate and transport for a geologically complex system like NERT study area, where small-scale preferential flow pathways cannot be fully and explicitly represented by the spatial discretization of the numerical regional model. Hence, for the current model, the dual porosity, mass transfer approach is used to represent back diffusion from the low conductivity UMCf.

7.3 Conceptual Perchlorate Mass Balance

A quarterly conceptual perchlorate mass removal was estimated for the model area for the period 2014-2018. There are three primary components of the perchlorate mass removal. The first component is perchlorate mass discharge to the Wash. This component is estimated based on an evaluation of perchlorate sampling for surface water at various locations in the Wash, paired with the corresponding instantaneous surface water flows from the nearest USGS gage station at the sampling time (see Section 3.2.1). The other two mass removal components within the model domain are the mass removal through extraction wells and the mass discharge through the storm drains, subdrains, and the mass leaving the model domain at the eastern boundary. The overall summary of the reported mass removal is provided in Table 9 and Figure 7-1.

7.3.1 Perchlorate Loading in the Las Vegas Wash

Quarterly perchlorate mass discharge to the Wash was estimated for each of the stream reaches for the period 2014-2018, as described in Section 3.2.1 and Table 1. The average total perchlorate mass discharge, estimated from groundwater discharge to the

Wash within the model boundary for the period 2014-2018, ranges from 50 pounds per day (lbs/d) (in 2018Q2) to 89 lbs/d (in 2015Q4), with an average of 75 lbs/d across the simulation period. Annual averages were 73 lbs/d in 2014, 79 lbs/d in 2015, 83 lbs/d in 2016, 76 lbs/d in 2017 and 62 lbs/d in 2018 (Table 9).

7.3.2 Perchlorate Mass Removal through the Extraction Wells

Perchlorate mass removed via the NERT extraction well fields (IWF, AWF and SWF) has been estimated using the monthly mass removal values previously reported within the annual remedial performance reports (Ramboll Environ 2014, 2015, 2016b; Ramboll 2017, 2018c, 2019b). Monthly values were aggregated to quarterly averages across the simulation period.

The average mass removal for the IWF was 598 lbs/d, with a minimum of 384 lbs/d in 2018Q3, and a maximum of 899 lbs/d in 2014Q1. For the AWF, the average across the simulation period was 470 lbs/d, with a minimum of 401 lbs/d in 2016Q2, and a maximum of 562 lbs/d in 2015Q2. For the SWF, the average across the simulation period was 67 lbs/d, with a minimum of 47 lbs/d in 2014Q2, and a maximum of 87 lbs/d in 2017Q4. The cumulative average mass removal from all well fields across the simulation period is approximately 1,140 lbs/d, with an average of 52% removal from the IWF, 42% removal from the AWF, and 6% removal from the SWF (Table 9). For the AP wells, measurements for perchlorate mass removal began in 2016Q4, with an average of 85 lbs/d. The minimum removal was 28 lbs/d during 2016Q4 and the maximum of 113 lbs/d occurred during 2017Q4.

For the Endeavour wells, monthly data was compiled from the semi-annual reports and then aggregated as quarterly averages, as was done for the flow rates described in Section 2.4. The average perchlorate removal was approximately 1,120 lbs/day for the simulation period, with a minimum of 910 lbs/d in 2018Q3 and a maximum of 1,390 lbs/d in 2014Q2 (AMPAC 2014, 2015a,b; Endeavour 2016a,b, 2017a,b, 2018a,b, 2019a).

Perchlorate removal via OSSM and TIMET wells are not reported in the monitoring reports; however, it is expected that these wells remove a negligible amount of perchlorate from the system. Since TIMET and OSSM are not treating extracted water for perchlorate, any removed perchlorate mass will be injected back in the system via their respective recharge trenches and therefore the mass removed from each of the TIMET and OSSM systems has been assumed to be 0 lbs/d for the conceptual summary (Table 9).

7.3.3 Perchlorate Discharge through the Athens Drainage Channel

Perchlorate discharge from the AMPAC plume through the Athens Drainage Channel (ADC-Main) is directly reported in the semi-annual performance monitoring reports provided by AMPAC and Endeavour (AMPAC 2014, 2015a,b; Endeavour 2016a,b, 2017a,b, 2018a,b, 2019a). The average perchlorate discharge via ADC was 14 lbs/d across the simulation period. The minimum removal of 8 lbs/d occurred during 2014Q4 and the maximum of 22 lbs/d was during 2017Q4 (Table 9).

As described in Section 3.4, the flow of the ADC channel was applied as focused recharge near the Wash as shown on Figures C1 to C5 (Appendix C). Using the similar approach, the perchlorate discharge from the channel is also applied near the Wash as concentrations in the recharge package. The total perchlorate discharge for each quarter is divided by the recharge area and has been applied as concentrations in the recharge package. Overall, the net perchlorate mass removed via the ADC channel is negligible.

7.3.4 Perchlorate Discharge through Weston Hills and Tuscany Sub-Drains

As described in Section 3.4, groundwater discharge through the Golf Course subdrains near the Wash is approximately 50 gpm (9,600 cfd). The perchlorate concentration in the outflow from these subdrains has been sampled monthly starting in October 2018 and continuing through July 2019 (Ramboll 2018b). The average concentration from these samples was approximately 2.34 mg/L. This results in a perchlorate mass removal of approximately 1.4 lbs/day, which was summarized as the estimated discharge for each quarter within the simulation period (Table 9). Using the same approach used for the ADC channel, the perchlorate mass removed via sub-drains is injected back in the model (at the discharge location near the Wash) using the recharge package. Hence, the net perchlorate mass removed via sub-drains in the simulation period is negligible.

For the Weston Hills subdrains, a surface discharge location has not been identified. It is possible that the discharge in these subdrains is included in the Tuscany sub-drains discharge location.

7.3.5 Perchlorate Discharge through the Eastern Boundary

The groundwater discharge leaving the model domain at the eastern boundary has been estimated as 25,000 cfd (see Section 3.6). There are two wells near this location, one that is north of the Wash (WMW3.5N) and one that is south of the Wash (WMW3.5S). Perchlorate sampling data is available at both locations starting in 2015 and continuing through 2018. For 2014, we have estimated the perchlorate concentrations using the respective quarterly averages for both locations, from 2015 through 2018. Quarterly average perchlorate concentrations were then multiplied by the estimated groundwater discharge rate to provide an estimate of perchlorate mass discharge through the eastern boundary, leaving the model domain. Quarterly averages are presented in Table 9 and the annual averages are 2.2 lbs/d for years 2014, 2016 and 2017. In 2015, the annual average was slightly higher, at 2.4 lbs/d and in 2018 it was slightly lower, at 2.1 lbs/d.

8. TRANSPORT MODEL DEVELOPMENT

The approach for the perchlorate fate and transport and the conceptual mass balance presented in Section 7 was used as the basis for the development of the transport model. The key transport model components are described in the following sections.

8.1 Transport Model Code

The transport model was developed using the MT3D-USGS, which is an updated version of the groundwater solute transport code MT3DMS. This code includes refined transport modeling capabilities to accommodate flow terms calculated by MODFLOW packages that were previously unsupported by MT3DMS (Bedekar et al. 2016). In particular, the MT3D-USGS code is able to simulate perchlorate fluxes to the Wash resulting from groundwater discharge. The MT3D-USGS also includes the capability to route a solute through dry cells that may occur in the Newton-Raphson formulation of MODFLOW (that is, MODFLOW-NWT) (Bedekar et al. 2016). The latest code received via email from the author is included in Appendix J.

8.2 Transport Model Simulation Period

Consistent with the flow model, the transport model simulates 1825.5 days over the period 2014-2018. The first half-day of the simulation is a steady-state stress period, followed by 1,825 days of transient simulation with a stress period for each quarter from 2014 to 2018.

8.3 Transport Model Calibration Objectives

The transport model was calibrated using a trial-and-error approach. The model calibration objectives were as follows:

- To match the simulated mass removal from the extraction wells within 15% of the observed mass removal estimates, as given in Table 9.
- To match simulated perchlorate loading from the groundwater discharge to the Wash for each stream reach.
- To obtain a mass balance error of less than 1%.
- To match perchlorate concentration targets with a residual standard deviation to range ratio of less than 10%.
- To match the simulated perchlorate plume with the observed plume for Q2-2018.

8.4 Transport Model Components

The key transport model parameters are described in the following sub-sections.

8.4.1 Dual Domain

As described in Section 7.2, the porous medium is simulated as two distinct domains, a mobile domain where groundwater flow can occur and an immobile domain where no flow can occur, and transport can only take place through mass transfer with the mobile domain. Instead of a single "effective" porosity for each model cell, two porosities are used to characterize the porous medium: one for the mobile domain and the other for the immobile domain.

A summary of the total and immobile porosity values assigned to different layers are presented in Table 10. The total porosity values are consistent with the values used for the perchlorate mass estimates, as described in Attachment A of the 2019 Annual Performance

Report (Ramboll 2019b). The mobile porosities are kept equal to the specific yield for each layer as presented in Table 4.

8.4.2 Perchlorate Calibration Targets

For the transport model, perchlorate concentration data measured in wells were used as model calibration targets. Quarterly average groundwater perchlorate concentration data for 2014-2018 were compiled from the NERT Project Database of analytical data. There sources of data include:

- Data collected by NERT:
	- − Data from the Groundwater Monitoring Program (Ramboll Environ 2014, 2015, 2016b; Ramboll 2017, 2018c, 2019b);
	- − Data from RI Implementation Phase 1, RI Implementation Phase 2, RI Implementation Phase 2 Parcel AB, and RI Implementation Phase 3;
	- − Data from the Unit Buildings 4 and 5 Investigation (Tetra Tech 2019c); and
	- Data from treatability studies;
		- o ZVI-Enhanced In-Situ Groundwater Treatability Study (Ramboll 2018a);
		- o AP Area Treatability Study (Tetra Tech 2018h);
		- o Galleria Drive Bioremediation Treatability Study (Tetra Tech 2018l);
		- o Groundwater Bioremediation Study (Tetra Tech 2016);
		- o In Situ Chromium Treatability Study (Tetra Tech 2018e);
		- o Las Vegas Wash Bioremediation Pilot Study (Tetra Tech 2017b, 2018j);
		- o Seep Well Field Area Bioremediation Treatability Study (Tetra Tech 2019b);
		- o Soil Flushing Treatability Study (Tetra Tech 2017a); and
		- o Vacuum Enhanced Recovery Treatability Study (Tetra Tech 2018g);
- Data received from SNWA, AMPAC/Endeavour, and BMI complex parties for use in the Annual Remedial Performance Reports for Chromium and Perchlorate for the NERT Site (Ramboll Environ 2014, 2015, 2016b; Ramboll 2017, 2018c, 2019b); and
- Data provided by AECOM from the Phase 1 Groundwater Investigation.

Groundwater perchlorate concentration data from monitoring wells and artesian wells were considered as targets. Samples marked as field duplicates were not included. In total, 926 wells and 6,298 perchlorate concentration measurements were considered for the model simulation period. The measurements were averaged on a quarterly basis in order to be compared to the simulated quarterly stress periods. Altogether, 4,548 quarterly perchlorate concentration targets were used for model calibration. Weights were applied to targets using the same method as for the groundwater elevation targets as described in Section 5.2.

Transport model targets used for model calibration are tabulated in Appendix H and are shown on Figure 8-1 and Figure 8-2. Figure 8-1 represents the location of calibration target in 2014. The calibration target locations for 2018 used for the model are shown on Figure 8- 2.

In addition, perchlorate loading in the Wash, mass removal through extraction wells, and the perchlorate mass discharge from the subdrains were used as qualitative calibration targets.

8.4.3 Perchlorate Initial Conditions

There are two perchlorate plumes within the model extent: the NERT plume and the AMPAC plume. For model Layers 1 through 3, the perchlorate plume presented in 2014 Annual Performance Report (Ramboll Environ 2014) was used to assign the initial concentration as shown on Figures 8-3, 8-4, and 8-5. However, the extent of the 2014 plume presented in the performance report is limited to the NERT site, the off-site RI areas, and the shallow AMPAC plume. In the source area for the AMPAC plume, the plume was interpolated based on perchlorate data presented in BCA Semi‐Annual/Annual Monitoring and Performance Report (AMPAC 2015b). For the Eastside Area, the initial concentrations were assigned based on 2018 plume in the 2018 Annual Performance Report (Ramboll 2018c). For Layers 4 through 10, due to limited availability of perchlorate data, initial perchlorate concentrations from the 2018 mass estimate were used and are presented on Figures 8-6 through 8-12.

Initially, equal perchlorate concentrations were assigned in the mobile and immobile zone for each grid cell in each model layer. The concentrations in the immobile zone was adjusted during calibration. The perchlorate mass present in the immobile domain is expected to serve as the continuous source of perchlorate in the model domain. Immobile concentrations under the Wash gravels in Layer 2 were increased to three times the mobile concentrations during calibration. The UMCf-cg underneath the NERT site is expected to have higher mass in the immobile zone. Hence, in Layer 2 the concentrations in the immobile zone in the UMCf-cg were increased to seven times the mobile concentrations. Immobile concentrations within the paleochannels were adjusted during calibration and were set to 1.5 times the mobile concentrations in Layer 1. The concentrations in the immobile domain were also adjusted in Layer 7 near AMEW extraction wells for calibrating the mass removal rates.

8.4.4 Transport Parameters

A summary of the transport parameter values assigned to different geologic units in the model are shown in Table 10 and are described below:

Dispersivity

Due to the impracticability of measuring dispersion in the field, dispersivity values are often estimated based on plume length or distance to receptors. For the NERT plume, the location of the leading edge of the plume is unknown due to Las Vegas Wash, which intercepts the plume. For the modeling purposes, the plume length of approximately 16,000 ft is assumed which is equal to the distance of the unit buildings from the Wash.

For the Phase 6 Model, the longitudinal dispersivity was estimated based on a formula developed by using a weighted best fit of field data (Xu and Eckstein 1995). This equation is provided below:

$$
a_x = 0.83[\log_{10}(L_p)]^{2.414}
$$

where $a_x =$ longitudinal dispersivity and $L_p =$ plume length in meters.

Based on the above equation, the estimated longitudinal dispersivity is around 64 ft. A uniform value of 60 ft for the dispersivity is used throughout the model domain. The

transverse dispersivity was assumed to be 10% of the longitudinal dispersivity, and the vertical dispersivity was assumed to be 1% of the longitudinal dispersivity.

Diffusion Coefficient

In an aqueous (water) solution, typical diffusion coefficients are in the range of 9.3 X 10^{-4} $ft²/d$ to 9.3 X10⁻⁵ $ft²/d$. As a result, diffusion in liquids is very slow over everyday length scales and is almost always dominated by advection. However, as stated in Section 7.1, in a low conductivity geologic material like UMCf-fg, diffusion can be significant. Hence, a relatively higher value of 1.53 X 10⁻³ ft²/day has been used in the model. This value is consistent with the value used in the University of Nevada, Las Vegas (UNLV) transport model (UNLV 2003).

Mass Transfer Coefficient

Perchlorate mass transfer from mobile to immobile zones occurs when there is a concentration gradient between the two zones. The magnitude of the exchange between the mobile and immobile domains is controlled by the mass transfer coefficient. As the mass transfer coefficient increases, the exchange between the mobile and immobile domains becomes increasingly fast and the dual-domain model functions more and more like a singledomain model whose porosity approaches the total porosity of the porous medium. On the other hand, as the mass transfer rate approaches zero, the dual-domain model becomes equivalent to a single-domain model with a porosity approaching the porosity of the mobile domain. For the current modeling work, a uniform mass transfer rate of 2.25 X 10⁻⁴ per day (1/d) has been used for alluvium, paleochannels, Wash gravels, and UMCf-cg. A higher value of 6.85 X 10⁻⁴ 1/d is used for transitional unit and the UMCf-fg.

8.4.5 Stream Boundary for Perchlorate Mass Loading

The MT3D-USGS can simulate solute mass exchange between surface water bodies that are connected to the groundwater using the Streamflow Transport (SFT) Package. The model simulates solute concentrations in stream reaches, where stream reaches are defined in the SFR package as described in Section 5.4.2. The SFT routes mass through stream networks and accounts for convergent flows, groundwater/surface water exchange, precipitation, and evaporation to and from stream surfaces, and overland runoff.

One of the inputs to the SFT package is the initial surface water concentrations of solute in each of the stream reaches. For this evaluation, the available surface water perchlorate data for 2014 was analyzed and interpolated within the Wash to get the perchlorate concentrations in each stream reach. These concentrations are then added into the input SFT file before running the final model.

One objective of the modeling is to estimate the mass flux of perchlorate from groundwater to the Wash, but it is not necessary to simulate the spatial distribution of chemical concentrations within the Wash. Hence, a uniform stream-dispersion coefficient of 9.3 X 106 ft²/d was applied to each stream reach in SFT package, which is a reasonable assumption for a small stream.

8.5 Transport Model Solver

To ensure a correct solution and to minimize the global mass balance error, the finite difference solution scheme was used for the transport model (Zheng 2010). The generalized conjugate gradient (GCG) solver was used with concentration change criteria of 0.001 milligrams per cubic feet (mg/ft³) and relaxation parameter as 1. An initial time step of 0.05 days was used for simulation.

9. TRANSPORT MODEL CALIBRATION

As described in Section 8.2, the transport model was run for 1,825.5 days to cover the period of 2014-2018.

9.1 Model Calibration

The transport model was manually calibrated to match the conceptual mass balance components for the simulation period (2014-2018). The mass transfer rate and the concentrations in the immobile zone for shallow layers were adjusted to get an acceptable match to measured mass removal rates in the extraction wells and the perchlorate loading in the Wash. The overall solute mass balance error was kept at less than 1%.

Table 11 provides a summary of target residual statistics for the 4,548 observation wells used to calibrate the Phase 6 transport model. The overall average target residual in the calibrated model is -0.57 mg/L. A negative residual value indicates that the simulated concentrations are higher than the observed values. There is a RMSE of 120.5 mg/L with the range of observation of 6,600 mg/L. The residual standard deviation of the model targets is 120.5 mg/L. A model is considered well calibrated when the ratio of the residual standard deviation to the range of observed values is less than 10 percent (Hill and Tiedeman 2007). For the Phase 6 transport model, this ratio is about 1.8% percent. The observed versus simulated concentration difference at the target locations is given in Appendix I. Residuals for each target location are shown in Table I-1 (Appendix I).

The calibration plots for each well show that there is a good fit at the vast majority of targets (Table I-1, Appendix I). Table I-1 shows the distribution of targets in the Shallow, Middle, and Deep WBZs. The match between simulated and observed concentrations appears reasonable except for results from AP Area and Soil Flushing treatability studies (Tetra Tech 2018h). The model is not expected to match these results accurately since soil flushing was not simulated in the model. Hence, for the treatability study wells, the model calibration was only focused on the baseline data points.

Figure 9-1 shows a scatter plot characterizing the match between modeled and observed groundwater perchlorate concentrations at wells used as calibration targets. The plots illustrate that there is generally good agreement between modeled and observed values, with most points falling close to the 1:1 correlation line. As stated above, for the treatability study wells, the calibration was refined to match baseline data only as shown on individual wells calibration plots in Appendix I. The "goodness-of-fit" $R²$ value is 0.88, demonstrating an acceptable fit to the observed concentrations.

9.2 Sensitivity Analysis of Transport Model

A sensitivity analysis was conducted to evaluate the effect of transport model input parameter uncertainty on significant model outputs. The model input parameters evaluated included: 1) dispersivity, 2) mass transfer rate, and 3) porosity. The sensitivity analysis was performed by changing each of these model input parameters by a factor of 0.75 and 1.25, running the model, and recording the outputs of interest.

The model outputs evaluated for the sensitivity analysis are: 1) calibration statistics based on simulated perchlorate concentrations at target locations, 2) the RMSE for the total perchlorate mass loading in the wash, and 3) the RMSE for the combined mass removal from the various extraction systems. The results of the sensitivity analysis for these outputs are shown in Table 12 and described in further detail below.

Dispersivity

The dispersivity value used in the calibrated model was 60 ft. When the value was reduced by 25%, the target residual statistics for majority of model outputs decreased, except for RMSE and sum of squared residuals of perchlorate in the groundwater, which has increased slightly. The model also has higher mass balance error with the lower dispersivity value. When the dispersivity value was increased by 25%, the target residual statistics for all model outputs increased except for the RMSE for mass removal in the extraction wells.

Mass Transfer Rate

Calibrated values for the mass transfer rate ranged from 2.25 X 10⁻⁴ to 6.85 X 10⁻⁴ ft²/d, depending on the location. By decreasing the mass transfer rate by 25%, the majority of the model output statistics increased. With an increased mass transfer rate by 25%, the statistics are general higher for all model outputs except for the RMSE for mass removal in the extraction wells, which decreased slightly.

Porosity

The calibrated values for total porosity ranged from 0.37 to 0.54, depending on the location within the model. By decreasing the porosity (both mobile and immobile) by 25%, all model output statistics has increased except for the RM for perchlorate in groundwater. By increasing the porosity (both mobile and immobile) by 25%, the model output statistics increased significantly except for the RMSE for perchlorate loading in the Wash, which decreased slightly.

9.3 Transport Model Results

The quarterly estimates of observed versus simulated mass removal rates for the model simulation period are presented on Figure 9-2. The calibration of the transport model is generally good based on a comparison of various simulated and conceptual mass balance components, mass removal at extraction wells, loading in the Wash, and the discharge in ADC Channel (Table 13).

Perchlorate Mass Loading in the Wash

The modeled versus observed perchlorate loading in the Wash at the stream gage stations is presented on Figures 9-3 through 9-7. The modeled estimates are comparable to the conceptual loading estimates at each station. However, the modeled estimated of loading in the initial time step of the model is generally lower than the conceptual estimates. The combined quarterly perchlorate loading in the Wash for the simulation period is also given in Table 13.

Perchlorate Mass Discharge in Athens Drainage Channel

The modeled versus observed perchlorate discharge in Endeavour's ADC channel is evaluated and is presented on Figure 9-8. The modeled discharge values are consistently higher by a few lbs/d as compared to the measured discharge values. The modeled discharge values are presented in Table 13.

Mass Removal Through Extraction Systems

The modeled perchlorate mass removal for each well field is evaluated and is presented in Table 13. The modeled mass removal is further compared to the measured mass removal for combined NERT and Endeavour's well field, and is presented on Figure 9-9 and Figure 9- 10, respectively. The mass removal by AP wells is combined with IWF mass removal for this evaluation (Figure 9-9).

The simulated mass removal rates at AWF are slightly lower than the observed mass removal in early part of model simulation. At the SWF, the simulated mass removal rates are estimated slightly higher than the observed rates. The simulated mass removal rates at other well fields are very similar to the measured mass removal (Table 13).

Mass Discharge at the Eastern Model Boundary

The simulated groundwater perchlorate discharge leaving the model domain at the eastern boundary is approximately 2-3 lbs/d during the model simulation period (Table 13). This has been estimated based on the total simulated flow in the outflow boundary at the eastern model boundary (Figures 5-11 and 5-12) and the average simulated perchlorate concentrations in wells WMW3.5N and WMW3.5S.

Simulated Perchlorate Plume for 2018

The simulated groundwater perchlorate plume for 2018 is compared with the observed perchlorate plume presented in 2018 Annual Performance Report (Ramboll 2018c). The comparative plume is shown on Figure 9-11. The simulated plume configuration reasonably matches the observed plume.

10. CONCLUSIONS

The Phase 6 Model simulates groundwater flow and perchlorate transport for the period 2014-2018. As part of the modeling effort, a detailed quarterly conceptual water balance and perchlorate mass balance was developed that was used as the basis for estimating the inflows and outflows of groundwater and perchlorate in the model.

The three-dimensional (3D) geological model developed to support the Phase 5 groundwater model was updated with borings drilled as part of the Phase 2 and 3 RI, as well as borings associated with treatability studies, including the Seep Well Field Area Bioremediation Treatability Study, Galleria Drive Bioremediation Treatability Study, and Las Vegas Wash Bioremediation Pilot Study. The boring logs for the RI Phase 2 off-site borings and the RI Phase 3 borings were still draft at the time the geological model was developed. There may be minor changes to these contact elevations in future versions of the model. Data collected as part of investigations by other parties has been incorporated in the model as calibration targets or for the evaluation of conceptual water balance.

The capability of simulating perchlorate fate and transport component was added to the Phase 6 Model based on the MT3D-USGS code using the dual-domain approach. The dual-domain approach was found to appropriately represent the effects of matrix diffusion. The Phase 6 Model was found to generally be able to accurately estimate the perchlorate loading in the Wash and the mass removal by the existing extraction systems.

The Phase 6 Model is expected to be further revised based on data collected as part of the OU-3 RI, recent investigations in the Downgradient Study Area, and ongoing treatability studies. In addition, the next version of the model will include the simulation of other major NERT COPCs, such as chlorate, chloroform, and hexavalent chromium. In order to simulate in-situ treatment of NERT COPCs, the next version of the model will also incorporate data from treatability studies concerning the feasibility of in-situ treatment in different areas of the NERT RI Study Area.

11. REFERENCES

- AECOM. 2019. Supplemental Surface Water Investigation Technical Memorandum; NERT Remedial Investigation - Downgradient Study Area; Nevada Environmental Response Trust Site; Henderson, Nevada. April 17.
- AMPAC (American Pacific Corporation). 2007. Report on the One Year Performance and Operations Optimization Program of the Endeavour In Situ Bioremediation (ISB) System for Perchlorate in Groundwater at the Athens Rd/Galleria Dr Area in Henderson, Nevada. August 23.
- AMPAC. 2012. BCA Semi-Annual/Annual Performance and BWPC UIC Monitoring Program; January 1 – June 30, 2012; Perchlorate In Situ Bioremediation System; American Pacific Corporation; Henderson, Nevada. August 14.
- AMPAC. 2014. BCA Semi-Annual/Annual Monitoring and Performance Report; January 1 to June 30, 2014; Perchlorate Bioremediation System; American Pacific Corporation; Henderson, Nevada. August 14.
- AMPAC. 2015a. BCA Semi-Annual/Annual Monitoring and Performance Report; July 1 to December 31, 2014; Perchlorate Bioremediation System; American Pacific Corporation; Henderson, Nevada. February 20.
- AMPAC. 2015b. BCA Semi-Annual/Annual Monitoring and Performance Report; January 1 to June 30, 2015; Perchlorate Bioremediation System; American Pacific Corporation; Henderson, Nevada. August 18.
- Bedekar, V., E.D. Morway, C.D. Langevin, and M. Tonkin. 2016. MT3D-USGS version 1.0.0: Groundwater Solute Transport Simulator for MODFLOW: U.S. Geological Survey Software Release. September 30. http://dx.doi.org/10.5066/F75T3HKD
- Bell J. and E. Smith. 1980. Geologic Map of the Henderson Quadrangle. Nevada Bureau of Mines and Geology, University of Nevada, Reno.
- Bingler. 1977. Geological Map of the Las Vegas SE Quadrangle. Nevada Bureau of Mines and Geology, University of Nevada, Reno.
- BRC (Basic Remediation Company). 2007. Conceptual Site Model Proposed CAMU Site, Henderson, Nevada, Submitted to Nevada Department of Environmental Protection, February 16.
- BRC. 2009. Summary Report for Updated Groundwater Flow Model Calibration, BMI Upper and Lower Ponds Area, Submitted to NDEP. November 2.
- CCDAQEM (Clark County Department of Air Quality and Environmental Management). 2009. Clark County 208, Area-Wide Water Quality Management Plan. March.
- Converse (Converse Consultants). 1998a. Subdrain Design Recommendations; Palm City Residential Development; Lake Mead Boulevard; Henderson, Nevada. Submitted to Rhodes Homes. February 6.
- Converse. 1998b. Infiltration Subdrain Design Recommendations; Palm City Residential Development; Lake Mead Drive and Future Sunset Road; Henderson, Nevada. Submitted to Rhodes Homes. November 11.
- de maximis (de maximis, inc.). 2014. Semi-Annual Operations Report; Groundwater Treatment System; Henderson, Nevada; January – June 2014. August 15.
- de maximis. 2015a. Semi-Annual Operations Report; Groundwater Treatment System; Henderson, Nevada; July – December 2014. February 15.
- de maximis. 2015b. Semi-Annual Operations Report; Groundwater Treatment System; Henderson, Nevada; January – June 2015. August 15.
- de maximis. 2016a. Semi-Annual Operations Report; Groundwater Treatment System; Henderson, Nevada; July – December 2015. February 15.
- de maximis. 2016b. Semi-Annual Operations Report; Groundwater Treatment System; Henderson, Nevada; January – June 2016. August 15.
- de maximis. 2017a. Semi-Annual Operations Report; Groundwater Treatment System; Henderson, Nevada; July – December 2016. February 15.
- de maximis. 2017b. Semi-Annual Operations Report; Groundwater Treatment System; Henderson, Nevada; January – June 2017. August 15.
- de maximis. 2018a. Semi-Annual Operations Report; Groundwater Treatment System; Henderson, Nevada; July – December 2017. February 15.
- de maximis. 2018b. Semi-Annual Operations Report; Groundwater Treatment System; Henderson, Nevada; January – June 2018. August 15.
- de maximis. 2019. Semi-Annual Operations Report; Groundwater Treatment System; Henderson, Nevada; July – December 2018. February 15.
- Donovan, D.J. and T. Katzer. 2000. Hydrogeologic Implications of Greater Ground-Water Recharge to Las Vegas Valley, Nevada. *Journal of the American Water Resources Association* 36(5):1133-1148.
- Donovan, D.J., T. Katzer, and K. Brothers. 2009. Review of Ground-Water Recharge Estimates in Nevada with an Analysis of Geologic Control on the Recharge Process, Nevada. Southern Nevada Water Authority. https://aquaterranova.files.wordpress.com/2014/09/2009-review-of-ground-waterrecharge-estimates.pdf
- Endeavour (Endeavour, LLC). 2016a. BISC Semi-Annual/Annual Monitoring and Performance Report; July 1 to December 31, 2015; Perchlorate Bioremediation System; Endeavour, LLC; Henderson, Nevada. February 15.
- Endeavour. 2016b. BISC Semi-Annual/Annual Monitoring and Performance Report; January 1 to June 30, 2016; Perchlorate Bioremediation System; Endeavour, LLC; Henderson, Nevada. August 11.
- Endeavour. 2017a. BISC Semi-Annual/Annual Monitoring and Performance Report; July 1 to December 31, 2016; Perchlorate Bioremediation System; Endeavour, LLC; Henderson, Nevada. February 14.
- Endeavour. 2017b. BISC Semi-Annual/Annual Monitoring and Performance Report, Rev.1; January 1 to June 30, 2017; Perchlorate Bioremediation System; Endeavour, LLC; Henderson, Nevada. August 14. Revised September 20.
- Endeavour. 2018a. BISC Semi-Annual Monitoring and Performance Report, Rev. 1; July 1 – December 31, 2017; Perchlorate Bioremediation System; Endeavour, LLC; Henderson, Nevada. February 14. Revised February 28.
- Endeavour. 2018b. BISC Semi-Annual/Annual Monitoring and Performance Report, Rev.1; January 1 to June 30, 2018; Perchlorate Bioremediation System; Endeavour, LLC; Henderson, Nevada. August 14. Revised September 14.
- Endeavour. 2019a. BISC Semi-Annual Monitoring and Performance Report, Rev.1; July 1 – December 31, 2018; Perchlorate Bioremediation System; Endeavour, LLC; Henderson, Nevada. February 15. Revised April 22.
- Endeavour. 2019b. Work Plan for Eastgate Drain Surface Water Mitigation System For Collection and Treatment at AGTS, Perchlorate Bioremediation System; Endeavour LLC; Henderson, Nevada. August 1.
- ENSR. 1999. National Pollution Discharge Elimination System, Permit Application Perchlorate Removal Action, Henderson, Nevada. ENSR Document No. 4020-011- 100. September.
- ENSR. 2005. Quarterly Performance Report; Perchlorate Recovery System; Kerr-McGee Chemical LLC, Henderson, Nevada. ENSR Document No. 04020-023. April.
- ENSR. 2007. Phase A Source Area Investigation Results Report, Tronox LLC Facility, Henderson, Nevada. September. NDEP approved November 30, 2007.
- ENVIRON (ENVIRON International Corporation). 2014a. Semi-Annual Remedial Performance Report for Chromium and Perchlorate. Nevada Environmental Response Trust Site; Henderson, Nevada; July 2013-December 2013. February 28.
- ENVIRON. 2014b. Annual Remedial Performance Report for Chromium and Perchlorate, Nevada Environmental Response Trust Site; Henderson, Nevada; July 2013-June 2014. October 31.
- Epstein, B. J., G.M. Pohll, J. Huntington and R.W.H. Carroll. 2010. Development and Uncertainty Analysis of an Empirical Recharge Prediction Model for Nevada's Desert Basins. *Journal of the Nevada Water Resources Association*. Summer 2010.
- GEI (GEI Consultants, Inc.). 2014. Titanium Metals Corporation, Henderson, Nevada Facility, Facility ID# H000537, Quarterly Status Report for the 4th Quarter of 2013. January 15.
- GEI. 2019. Groundwater Extraction and Treatment System Operations, Maintenance and Performance Report; Fourth Quarter 2018. February 28.
- Geosyntec (Geosyntec Consultants). 2010. Groundwater Flow Model, South of Warm Springs Study Area; Henderson, Nevada; Prepared for American Pacific Corporation. February 3.
- Geosyntec. 2013. Transient Calibration of Groundwater Flow Model South of Warm Springs Study Area Henderson, Nevada. June 28.
- Geosyntec. 2016. Groundwater Flow Model Update, Henderson, Nevada. September 13.
- Geosyntec. 2017. Shallow Zone Capture Assessment Revision 1, Endeavour, Henderson, Nevada. March 23.
- GES (Geotechnical & Environmental Services, Inc.). 2003. Geotechnical Investigation Demonstration Weir W-3.9 Replacement SNWA Contract No. 810-K, Las Vegas Wash Erosion Control Program, Clark County, Nevada, File No. 20031518E1. August 8.
- GES. 2007. Deep Background Investigation Report. November 19.
- Geovision (Geovision Geophysical Service). 2003a. Draft Interim Report, Geophysical Investigation; BMI Upper and Lower Ponds and Ditches; Clark County, Nevada. May 1.
- Geovision. 2003b. Preliminary Draft Report, Geophysical Investigation; Las Vegas Wash; Clark County, Nevada. May 1.
- Hanson, R.T., S.E. Boyce, Schmid, Wolfgang, J.D. Hughes, S.M. Mehl, S.A. Leake, Maddock, Thomas, III and R.G. Niswonger. 2014. One-Water Hydrologic Flow Model (MODFLOW-OWHM): U.S. Geological Survey Techniques and Methods 6– A51. http://dx.doi.org/10.3133/tm6A51.
- Hargis (Hargis+Associates, Inc.). 2008. Conceptual Site Model Former Montrose and Stauffer Facilities and Downgradient Areas to Las Vegas Wash, Henderson, Clark County, Nevada. July 25.
- Hargis. 2012. 2012 Comprehensive Groundwater Data Evaluation Report; Former Montrose and Stauffer Facilities and Current Olin Facility; Henderson, Nevada. August 16.
- Hill, M.C. and C.R. Tiedeman. 2007. *Effective Groundwater Model Calibration: With Analysis of Data, Sensitivities, Predictions, and Uncertainty*. John Wiley & Sons, Inc.
- Kerr-McGee. 2000. Semi-Annual Chromium Mitigation Performance Report, Kerr-McGee Chemical LLC, Henderson, Nevada, January 26.
- Kerr-McGee. 2001. Semi-Annual Performance Report; Chromium Mitigation Program. July 27.
- Kleinfelder. 2005. Hydrogeologic Investigation 2004 Phase III Drilling and Aquifer Testing; American Pacific Corporation, Henderson, Nevada. January 11.
- Konikow, L.F., G.Z. Hornberger, K.J. Halford and R.T. Hanson. 2009. Revised multi-node well (MNW2) package for MODFLOW ground-water flow model: U.S. Geological Survey Techniques and Methods 6–A30.
- LVWCC (Las Vegas Wash Coordination Committee). 2019. Year-End Report, 2018. April. https://www.lvwash.org/assets/pdf/resources_lvwcc_2018.pdf
- Maxey, G.B. and T.E. Eakin. 1949. Ground Water in White River Valley, White Pine, Nye, and Lincoln Counties, Nevada. State of Nevada, Office of the State Engineer.
- McGinley (McGinley & Associates). 2003. Las Vegas Wash Initial Perchlorate Modeling Report, Las Vegas Wash; Henderson, Nevada*;* Prepared for Nevada Division of Environmental Protection. October 20.
- McGinley. 2007. Athens Road Well Field Modeling Report, Near BMI Industrial Complex; Henderson, Nevada; Prepared for Nevada Division of Environmental Protection. June 30.
- NRC (National Research Council). 2013. *Alternatives for Managing the Nation's Complex Contaminated Groundwater Sites*. National Academies Press: Washington DC.
- NDEP (Nevada Division of Environmental Protection). 2003. Las Vegas Wash Initial Perchlorate Modeling Report, Las Vegas Wash; Henderson, Nevada. October 20.
- NDEP. 2009. Letter Re: BMI Plant Sites and Common Area Projects, Henderson, Nevada; Hydrogeologic and Lithologic Nomenclature Unification. January 6.
- NDEP. 2017. Response to: Phase 5 Transient Groundwater Flow Model, Response to Comments, Nevada Environmental Response Trust, Henderson, Nevada. June 26.
- Niswonger, R.G., Panday, Sorab, and M. Ibaraki. 2011. MODFLOW-NWT, A Newton formulation for MODFLOW-2005: U.S. Geological Survey Techniques and Methods 6– A37.
- Northgate (Northgate Environmental Management, Inc.). 2010. Capture Zone Evaluation Report; Tronox LLC; Henderson, Nevada. December 3.
- PRISM (PRISM Climate Group). 2013. Gridded Climate Data for the Contiguous USA. http://prism.oregonstate.edu.
- Ramboll Environ (Ramboll Environ US Corporation). 2014. Annual Remedial Performance Report for Chromium and Perchlorate, Nevada Environmental Response Trust Site; Henderson, Nevada; July 2013-June 2014. October 31. NDEP approved December 26.
- Ramboll Environ. 2015. Annual Remedial Performance Report for Chromium and Perchlorate, Nevada Environmental Response Trust Site; Henderson, Nevada; July 2014-June 2015. October 29. NDEP approved December 30.
- Ramboll Environ. 2016a. Phase 5 Transient Groundwater Flow Model Nevada Environmental Response Trust Site; Henderson, Nevada. November 15. NDEP approved June 26, 2017.
- Ramboll Environ. 2016b. Annual Remedial Performance Report for Chromium and Perchlorate, Nevada Environmental Response Trust Site; Henderson, Nevada; July 2015-June 2016. October 31. NDEP approved December 6.
- Ramboll Environ. 2016c. Technical Memorandum, Remedial Investigation Data Evaluation, Nevada Environmental Response Trust Site; Henderson, Nevada. May 2. NDEP approved July 13.
- Ramboll Environ. 2017. RI Study Area Mass Estimate and Expanded Performance Metrics Technical Approach, Nevada Environmental Response Trust Site; Henderson, Nevada. October 5. NDEP approved October 19, 2017.
- Ramboll (Ramboll US Corporation). 2017. Annual Remedial Performance Report for Chromium and Perchlorate, Nevada Environmental Response Trust Site; Henderson, Nevada; July 2016-June 2017. December 8. NDEP approved February 6, 2018.
- Ramboll. 2018a. Technical Memorandum, Treatability/Pilot Study Modification No. 5, Nevada Environmental Response Trust Site; Henderson, Nevada. November 5. NDEP approved November 14.
- Ramboll. 2018b. RI Phase 3 Modification No. 2, Nevada Environmental Response Trust Site; Henderson, Nevada. Ramboll Project no. 1690006943-036. April 10. NDEP approved April 12.
- Ramboll. 2018c. Annual Remedial Performance Report for Chromium and Perchlorate, Nevada Environmental Response Trust Site; Henderson, Nevada; July 2017-June 2018. November 9. NDEP approved January 18, 2019.
- Ramboll. 2019a. Barrier Wall Integrity Evaluation Report, Nevada Environmental Response Trust Site; Henderson, Nevada. August 30. Revised November 21.
- Ramboll. 2019b. Annual Remedial Performance Report for Chromium and Perchlorate, Nevada Environmental Response Trust Site; Henderson, Nevada; July 2018–June 2019. In progress.
- Sanford, W.E. and D.L. Selnick. 2012. Estimation of Evapotranspiration Across the Conterminous Unites States using a Regression with Climate and Land-Cover Data. *Journal of the American Water Resources Association* 49(1): 217-230.
- Tetra Tech (Tetra Tech, Inc.). 2016. Groundwater Bioremediation Treatability Study Results Report; Nevada Environmental Response Trust Site; Henderson, Nevada. November 25.
- Tetra Tech. 2017a. Soil Flushing Treatability Study Report; Nevada Environmental Response Trust Site; Henderson, Nevada. July 3.
- Tetra Tech. 2017b. Las Vegas Wash Bioremediation Pilot Study Work Plan; Nevada Environmental Response Trust Site; Henderson, Nevada. September 22.
- Tetra Tech. 2018a. Operation and Maintenance Summary January 2018 Weir Dewatering Treatment Plant. February 28.
- Tetra Tech. 2018b. Operation and Maintenance Summary February 2018 Weir Dewatering Treatment Plant. March 23.
- Tetra Tech. 2018c. Operation and Maintenance Summary March 2018 Weir Dewatering Treatment Plant. April 20.
- Tetra Tech. 2018d. Operation and Maintenance Summary April 2018 Weir Dewatering Treatment Plant. May 22.
- Tetra Tech. 2018e. In-Situ Chromium Treatability Study Results Report; Nevada Environmental Response Trust Site; Henderson, Nevada. June 22.
- Tetra Tech. 2018f. Operation and Maintenance Summary May 2018 Weir Dewatering Treatment Plant. June 22
- Tetra Tech. 2018g. Vacuum Enhanced Recovery Treatability Study Results Report; Nevada Environmental Response Trust Site; Henderson, Nevada. July 12.
- Tetra Tech. 2018h. AP Area Down and Up Flushing Treatability Study Results Report, Nevada Environmental Response Trust Site, Henderson, July 27.
- Tetra Tech. 2018i. Operation and Maintenance Summary June 2018 Weir Dewatering Treatment Plant. July 31.
- Tetra Tech. 2018j. Treatability/Pilot Study Modification No. 2 Las Vegas Wash Bioremediation Pilot Study; Nevada Environmental Response Trust Site; Henderson, Nevada. August 17.
- Tetra Tech. 2018k. Operation and Maintenance Summary July 2018 Weir Dewatering Treatment Plant. August 20.
- Tetra Tech. 2018l. Treatability/Pilot Study Modification No.3, Revision 1 Galleria Road Bioremediation Treatability Study; Nevada Environmental Response Trust Site; Henderson, Nevada. August 29.
- Tetra Tech. 2018m. Operation and Maintenance Summary August 2018 Weir Dewatering Treatment Plant. September 20.
- Tetra Tech. 2018n. Operation and Maintenance Summary September 2018 Weir Dewatering Treatment Plant. November 5.
- Tetra Tech. 2018o. Operation and Maintenance Summary October 2018 Weir Dewatering Treatment Plant. December 3.
- Tetra Tech. 2018p. Operation and Maintenance Summary November 2018 Weir Dewatering Treatment Plant. December 27.
- Tetra Tech. 2019a. Operation and Maintenance Summary December 2018 Weir Dewatering Treatment Plant. January 21.
- Tetra Tech. 2019b. Seep Well Field Area Bioremediation Treatability Study Results Report; Nevada Environmental Response Trust Site; Henderson, Nevada. August 13.
- Tetra Tech. 2019c. Unit 4 and 5 Buildings Investigation Source Area Characterization Report; Nevada Environmental Response Trust Site; Henderson, Nevada. August 29.
- TIMET (Titanium Metals Corporation Facility). 2008. Semiannual Groundwater Monitoring Report, 2nd Semester 2008; Titanium Metals Corporation Facility; Henderson, Nevada. October 31.
- Todd, D.K. 1980. *Groundwater Hydrology*, 2nd ed., New York: John Wiley & Sons.
- UNLV (University of Nevada, Las Vegas). 2003. Fate and Transport of Perchlorate in a Contaminated Site in the Las Vegas Valley. Prepared for USEPA.
- USGS (United States Geological Survey). 1989. Groundwater Conditions in Las Vegas Valley, Clark County, Nevada, Part 1, Hydrogeologic Framework; United States Geological Survey Water-Supply Paper 2320-A.
- USGS. 1996. Groundwater Conditions in Las Vegas Valley, Clark County, Nevada, Part 2, Hydrogeology and Simulation of Ground-Water Flow. United States Geological Survey Water-Supply Paper 2320-B.
- USGS. 1997. Preliminary Location and Geometry of the Las Vegas Valley Shear Zone Based on Gravity and Aeromagnetic Data. Open-File Report 97-441.
- USGS. 2004a. Groundwater Pumpage and Artificial Recharge Estimates for Calendar Year 2000 and Average Annual Natural Recharge and Interbasin Flow by Hydrographic Area, Nevada; United States Geological Survey Scientific Investigation Report 2004- 5239.
- USGS. 2004b. A New Streamflow-Routing (SFR1) Package to Simulate Stream-Aquifer Interaction with Modflow-2000; United States Geological Survey Open-File Report 2004-1042.
- USGS. 2005. Geologic and Geophysical Maps of the Las Vegas 30' X 60' Quadrangle, Clark and Nye Counties, Nevada, and Inyo County, California; United States Geological Survey - Pamphlet to accompany Scientific Investigations Map 2814.
- USGS. 2008. Quantifying Ground-Water and Surface-Water Discharge from Evapotranspiration Processes in 12 Hydrographic Areas of the Colorado Regional Ground-Water Flow System, Nevada, Utah, and Arizona; United States Geological Survey Scientific Investigation Report 2008–5116.
- Vector (Vector Engineering Inc.). 2011. Laboratory Services, Soil Bentonite, Backfill Mix Design Henderson Slurry Wall, Nevada. April 12.
- Xu, M. and Y. Eckstein. 1995. Use of weighted least-squares method in evaluation of the relationship between dispersivity and field scale. *Groundwater* 33(6): 905-90.
- Zheng, C. 2010. MT3DMS v5.3 Supplemental User's Guide, Department of Geological Sciences, University of Alabama, Tuscaloosa.

Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

TABLES

TABLE 1. ESTIMATED PERCHLORATE MASS DISCHARGE TO LAS VEGAS WASHPhase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

lbs/d = pounds per day

[#] = Mass discharge reported by AECOM, Supplemental Surface Water Investigation Technical Memorandum (2019).

a = Discharge data were not available for Las Vegas Wasteway from January 2016 through February 2017, and there was only one sample available for 1Q of 2017. The average of 1Q from 2015 and 2018 was used to estimate the loading for 1Q 2017 and then the average quarterly loadings from 2015 and 2017 were used to estimate the loadings for all quarters in 2016, respectively.

^b = Measurements were not available for 2014 through 2016 for the Duck Creek Confluence, Bostic Weir and Homestead Weir stations. Loading for these locations was estimated using the average distribution of loading for each sub-reach in 2017.

^c = Loading values at Pabco Road reflect the quarterly averages as they have been reported by Ramboll in the Semi-annual and Annual Remedial Performance Reports[.]

 d = The loading for 3Q 2017 was slightly higher at Bostic Weir than the loading estimated at Pabco Road. To avoid a negative loading value (-1 lb/d), this value was adjusted to zero.

 e = Loading data became available for the Bostic and Homestead gages at the end of June 2017. The average loading from 3Q-4Q 2017 was instead used to estimate the loading for 2Q of 2017. For 1Q of 2017 at these locations, the loading was estimated as the average of 4Q in 2017 and 2018, to reflect the similar seasonal conditions in the winter.

f = At the Three Kids location, perchlorate concentrations and streamflows were abnormally high between 4Q 2015 through 4Q 2016. During this time period, the average loading measured between Pabco Road and Northshore Road has been used in place of the loading between Pabco Road and Three Kids. The loading for Reach B was then distributed to the sub-reaches according to the average distribution in 2017 for loading between Pabco Road and Three Kids.

TABLE 2a. GROUNDWATER DISCHARGE TO LAS VEGAS WASH BASED ON CHEMISTRY AND FLOW DATA Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

a Perchlorate concentrations are calculated using well sampling data in the second quarter of each year. Wells are selected using a 300 foot buffer zone to the south of the Las Vegas Wash, except for Reach 1, where well PM *Net combined groundwater inflow rate rounded to two significant figures after calculating.*

Notes:

cfd = cubic feet per day

cfs = cubic feet per second

lbs/d = pounds per day

µg/l = micrograms per liter

¹ The average perchlorate loading at LV Wasteway Station for 2017 is 1.02 lbs/day. To get perchloate loading in Reach 1, the 1.9 lbs/d measured at the Duck Creek Confleunce is subtracted form the LV wasteway station.

TABLE 2b. ESTIMATED WATER BALANCE IN LAS VEGAS WASHPhase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

^b Annual evaporation rates calculated from stream area within each reach and average evaporation rate reported in Moreo, M.T. and A. Swancar. 2013. Evaporation from Lake Mead, Nevada and Arizona, March 2010 through Febru Scientific Investigations Report 2013–5229, 40 p., http://dx.doi.org/10.3133/sir20135229) ISSN 2328-0328 (online).

Notes:

cfs = cubic feet per second

-- = Not Applicable

NA= Not Available (Duck Creek Confluence, Bostic, and Homestead gage stations were installed in September 2016. Las Vegas Wasteway gage station has no data in 2016.)

AMPAC = American Pacific Corporation (now Endeavour, LLC)

COH = City of Henderson

NERT = Nevada Environmental Response Trust

TIMET = Titanium Metals Corporation

^a From Table 2a

TABLE 2c. GROUNDWATER EXTRACTION RATES Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

TABLE 2c. GROUNDWATER EXTRACTION RATES Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

TABLE 2c. GROUNDWATER EXTRACTION RATES Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

cfd = cubic feet per day

ft msl = feet above mean sea level

-- = Not available

AMPAC = American Pacific Corporation (now Endeavour, LLC)

NERT = Nevada Environmental Response Trust

OSSM = Olin Chlor-Alkali/Stauffer/Syngenta/Montrose

SNWA = Southern Nevada Water Authority

TIMET = Titanium Metals Corporation

^a NERT data obtained from GWETS Field Sheets.

 b OSSM data are obtained from OSSM monitoring and operations reports.

 c AMPAC data are compiled from AMPAC monitoring reports.

 d TIMET data are compiled from TIMET monitoring reports.

e. Weir dewatering data are compiled from dewatering water treatment operation and maintenance summaries.

TABLE 2d. CONCEPTUAL GROUNDWATER BALANCE SUMMARYPhase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

AWF = Athens Road Well Field

n Applied as 50% of total weir dewatering (see Table 2c).

^o Estimated net discharge (See Table 1).

^p Field Estimation.

Maxey, G.B. and T.E. Eakin. 1949. Ground Water in White River Valley, White Pine, Nye, and Lincoln Counties, Nevada. State of Nevada, Office of the State Engineer.

^a Boundary inflows are expected to be constant during the model simulation period. https://educ.backgroundary.html be rechended be constant during the model simulation period. https://educ.backgroundary.html basins, from

^c Based on USGS (2004).

form Site Development." Conference Paper 100, OpenSIUC, Southern Illinois University Carbondale.

from Site Development." Conference Paper 100, OpenSIUC, Southern Illinois University Carbondale.

The total inflow and outflow values have been rounded to three significant figures, after calculations.

-- \equiv Not Applicable b estimated based on watershed area and precipitation recharge (Table 2e).
The extending on a real of the state of the state of phreatophytic vegetation and evapotranspiration rate.

IWF = Interceptor Well Field \blacksquare Estimated as inflow rate plus precipitation rate minus evaporation rate.

SWF = Seep Well Field **Exercise 2 and Spreadsheet.** ^e Based on estimate of leakage from water mains. **i NERT Field Spreadsheet.**

AMPAC = American Pacific Corporation (now Endeavour, LLC) ^f Based on water usage and percent estimate to prevent salt buildup in the root zone. ^j RI Phase 3, Modification #2. NERT = Nevada Environmental Response Trust expected in undeveloped areas based on the resources below: k AMPAC/Endeavor Monitoring Reports.

OSSM = Olin Chlor-Alkali/Stauffer/Syngenta/Montrose sanford W. Eand D.I. Selpick 2013. Estimation of Evanotranspiration Across the santo laterly average pumping. Sanford, W. E. and D.L.Selnick. 2013. Estimation of Evapotranspiration Across the Equal to quarter and D.L.Selnick. 2013. Estimation of Evapotranspiration Across the Equal to quarter and Land-Cover Data. The Tritanium Meta

TABLE 2e. EVALUATION OF UNCERTAINTY IN CONCEPTUAL BOUNDARY INFLOWS Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site

Henderson, Nevada

Notes:

cfd = cubic feet per day

Donovan, D.J. and T. Katzer. 2000. Hydrogeologic Implications of Greater Ground-Water Recharge to Las Vegas Valley, Nevada. *Journal of the American Water Resources Association* 36(5): 1133-1148.

Epstein, B. J., G.M. Pohll, J. Huntington and R.W.H. Carroll. 2010. Development and Uncertainty Analysis of an Empirical Recharge Prediction Model for Nevada's Desert Basins. *Journal of the Nevada Water Resources Association*. Summer 2010.

Total boundary inflows were rounded to two significant figures, after calculations.

TABLE 3. MODEL LAYERS Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Henderson, Nevada

Notes:

 $ft = feet$

UMCf-fg = Fine-grained Upper Muddy Creek Formation

UMCf-cg = Coarse-grained Upper Muddy Creek Formation

xMCf = Transitional Muddy Creek Formation

W, C and E regions of the UMCf-cg correspond to model areas underlying the southwest

(AMPAC/Endeavour area), south central (upgradient NERT area), and southeast (BMI upper pond area) portions of the model domain, respectively.

TABLE 4. ESTIMATED VERSUS CALIBRATED HYDRAULIC CONDUCTIVITY AND STORAGE PROPERTIES Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

TABLE 4. ESTIMATED VERSUS CALIBRATED HYDRAULIC CONDUCTIVITY AND STORAGE PROPERTIES Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

 $ft = feet$

ft/day = feet per day

-- = Not Applicable

Qal = Quaternary Alluvium

UMCf = Upper Muddy Creek Formation

UMCf-cg = UMCf, coarse-grained

xMCf = Transitional Muddy Creek Formation

a Statistics derived from compiled historic and recent aquifer testing results (see Appendix D).

b Statistics derived from the vertical conductivity data in Appendix E4 of the Phase 5 Model (Ramboll Environ 2016), see Appendix D.

c Values for specific yield are assumed to be equal to the mobile porosity.

d Statistics derived from the storage parameter information in Appendix E5 of the

Phase 5 Model (Ramboll Environ 2016), updated in Appendix D.

TABLE 5. WEIR SPECIFICATIONSPhase 6 Groundwater Flow and Transport Model

Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

ft = feet

ft/ft = feet per foot

ft msl = feet above mean sea level

yr = year

-- = Not Applicable

NE = Not Evaluated (outside of the model boundary)

BSP = Bottom of Sheet Pile

EL = Elevation

SP = Sheet Pile

^a Contact elevation of Alluvium and Upper Muddy Creek Formation, as implemented in the model

Weir names are listed in order of upstream to downstream

TABLE 6. SIMULATED GROUNDWATER INFLOWS AND OUTFLOWS

Phase 6 Groundwater Flow and Transport Model

Nevada Environmental Response Trust Site

Henderson, Nevada

Notes:

cfd = cubic feet per day

-- = Not Applicable

AWF = Athens Road Well Field

IWF = Interceptor Well Field

SWF = Seep Well Field

AMPAC = American Pacific Corporation (now Endeavour, LLC)

NERT = Nevada Environmental Response Trust

OSSM = Olin Chlor-Alkali/Stauffer/Syngenta/Montrose

TIMET = Titanium Metals Corporation

The total inflow and outflow values have been rounded to three significant figures, after calculations.

TABLE 7. CALIBRATION STATISTICS FOR FLOW MODEL Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

 $ft = feet$ ft² = square feet

TABLE 8. SENSITIVITY ANALYSIS OF FLOW MODEL RESULTS Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

TABLE 8. SENSITIVITY ANALYSIS OF FLOW MODEL RESULTS Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

 $ft = feet$

ft² = square feet

ft/d = feet per day

cfd = cubic feet per day

cfs = cubic feet per second

LVW = Las Vegas Wash

RM = residual mean

RMSE = root mean square error

UMCf-fg = Fine-grained Upper Muddy Creek Formation

UMCf-cg = Coarse-grained Upper Muddy Creek Formation

Calibration statistics of calibrated model are highlighted

TABLE 9. CONCEPTUAL PERCHLORATE MASS REMOVAL SUMMARYPhase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

AWF = Athens Road Well Field

AMPAC = American Pacific Corporation (now Endeavour, LLC)

lbs/d = pounds per day a Based on the values presented in the Ramboll Annual Performance Reports

-- = Not Applicable b Perchlorate removal is not estimated

^c Based on the values and/or figures presented in the AMPAC/Endeavor Monitoring Reports

IWF = Interceptor Well Field d Based on the values reported by Tetra Tech in their monthly dewatering reports.

SWF = Seep Well Field entity and the Estimated using streamflow and perchlorate concentrations in the Wash.

f Estimated from data collected from November 2018 through July 2019 as part of RI Phase 3, Modification #2.

^g Estimated from average perchlorate concentrations measured at wells WMW3.5S and WMW3.5N. Data for 2014 only included WMW3.5S samples, so the quarterly averages for 2014 reflects the respective quarterly averages from 2015 through 2018.

NERT = Nevada Environmental Response Trust OSSM = Olin Chlor-Alkali/Stauffer/Syngenta/Montrose

TIMET = Titanium Metals Corporation

TABLE 10. TRANSPORT MODEL PARAMETERS Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

ft = feet

 ft^2/d = square feet per day

UMCf-cg = Coarse-grained Upper Muddy Creek Formation

UMCf-fg = Fine-grained Upper Muddy Creek Formation

xMCf = Transitional Muddy Creek Formation

The mobile porosities are same as specific yield presented in Table 4.

TABLE 11. CALIBRATION STATISTICS FOR TRANSPORT MODEL Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

mg/l = milligram per liter $[mg/l]^2$ = milligram per liter, squared

TABLE 12. SENSITIVITY ANALYSIS OF TRANSPORT MODEL RESULTSPhase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

 $ft = feet$

lbs/d = pounds per day

mg/l = milligrams per liter

 $[mg/l]^2$ = milligram per liter, squared

RM = residual mean

RMSE = root mean square error

Calibration statistics of calibrated model are highlighted

TABLE 13. MODELED PERCHLORATE MASS REMOVAL SUMMARY Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

lbs/d = pounds per day

AWF = Athens Road Well Field

IWF = Interceptor Well Field

SWF = Seep Well Field

AMPAC = American Pacific Corporation (now Endeavour, LLC)

NERT = Nevada Environmental Response Trust

OSSM = Olin Chlor-Alkali/Stauffer/Syngenta/Montrose

TIMET = Titanium Metals Corporation

Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

FIGURES

EXPLANATION

Phase 6 Model Extent

- ----- Geologic Faults
	- **NERT Property Boundary**
- Direction of Groundwater Flow

Primary Lithologies

-
-
-

H:\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 2-2 - Hydrology-Las Vegas Wash_RZ.mxd

-
-
-

- \overline{s} Surface Flow Capture Sump (Not operational since 2007)
- Slurry Wall
-
- &< Pumping Wells
	- NERT Recharge Trenches
- NERT / AMPAC (Endeavour) /

O TIMET / COH Outfall Location
	- Phase 6 Model Extent
	- NERT Property Boundary

AMPAC: American Pacific Corporation, now Endeavour LLC COH: City of Henderson NERT: Nevada Environmental Response Trust **OREAD SERVIS CORPORT MERT AMPAC (Entitant Corporation Phase 6 Model Extra)

NERT / AMPAC (Entitant Phase 6 Model Extra)

Phase 6 Model Extra

NERT Property Bou

NOTE:

AMPAC: American Pacific Corporation

NERT: Nevada Env**

MODEL AREA FEATURES: REMEDIATION SYSTEM Nevada Environmental Response Trust Site

Henderson, Nevada

EXPLANATION

NOTE:

ramboll\projects\NERT\Modeling\Phase_6\figures\LVW_station_discharge.py

ramboll\projects\NERT\Modeling\Phase_6\figures\LVW_station_discharge.py

ramboll\projects\NERT\Modeling\Phase_6\figures\LVW_station_discharge.py

MODEL AREA FEATURES: UNLINED PONDS

Nevada Environmental Response Trust Site

Path: H:\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 3-13 - Model Area Features - Unlined Ponds.mxd

0 0.5

EXPLANATION

- Phase 6 Model Extent
- **NERT Property Boundary**
- Bird Viewing Preserve, unlined
- Pond 13, unlined
- Las Vegas Wash
- Gravel Pit Pond

Path: H:\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 3-16 - Conceptual Boundary Inflows.mxd

Note: Cross sections are not drawn at the same scale. Elevations are in feet above mean sea level. XY coordinates are in Nevada State Plane East (feet).

NERT = Nevada Environmental Response Trust Site UMCF-fg = Fine-grained Upper Muddy Creek Formation UMCF-cg = Coarse-grained Upper Muddy Creek Formation

EXPLANATION

- NERT Property Boundary
- \otimes Borehole Intersecting UMCf-cg
- Borehole Data Below 100 feet \bullet
	- Fraction Coarse-Grained [a]

UMCf-cg Thickness (feet)

Note:

[a] Ratio of thickness of UMCf-cg to total UMCf within the vertical extent of the model.

UMCf = Upper Muddy Creek Formation UMCf-cg = Coarse-grained UMCf

- NERT Property Boundary
- Borehole Intersecting UMCf-fg and cg
- Borehole Data Below 100 feet
- Fraction Fine-Grained [a]

[a] Ratio of thickness of UMCf-fg to total UMCf within the vertical extent of the model.

UMCf = Upper Muddy Creek Formation UMCf-cg = Coarse-grained UMCf

1 Mile

-
-

UMCf-cg = Coarse-grained Upper Muddy

Model Hydraulic Conductivity in Layer 2 Nevada Environmental Response Trust Site Henderson, NV

EXPLANATION

- Phase 6 Model Extent
- NERT Property Boundary

Hydraulic Conductivity (ft/d)

Note: Horizontal hydraulic conductivity values shown are in feet per day.

0.66 Horse Spring Formation

UMCf-cg = Coarse-grained Upper Muddy Creek Formation xMCf = Transitional Muddy Creek Formation

RAMBCLL

Path: H:\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 5-3 - Model Hydraulic Conductivity in Layer 3.mxd

EXPLANATION

Phase 6 Model Extent

NERT Property Boundary

Hydraulic Conductivity (ft/d)

Note: Horizontal hydraulic conductivity values shown are in feet per day.

0.66 0.72 1 10 Horse Spring Formation UMCf-fg UMCf-cg

UMCf-fg = Fine-grained Upper Muddy Creek Formation UMCf-cg = Coarse-grained Upper Muddy Creek Formation

Model Hydraulic Conductivity in Layer 6 to 7 Nevada Environmental Response Trust Site

Path: H:\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 5-5 - Model Hydraulic Conductivity in Layer 6 to 7.mxd

UMCf-fg = Fine-grained Upper Muddy Creek Formation UMCf-cg = Coarse-grained Upper Muddy Creek Formation

Model Hydraulic Conductivity in Layer 8 Nevada Environmental Response Trust Site Henderson, NV

Path: H:\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 5-6 - Model Hydraulic Conductivity in Layer 8.mxd

EXPLANATION

 $\overline{}$ 10

Phase 6 Model Extent NERT Property Boundary **Hydraulic Conductivity (ft/d)** 0.66 Horse Spring Formation 0.72 — UMCf-fg 1 \blacksquare 1.2 UMCf-cg

Note: Horizontal hydraulic conductivity values shown are in feet per day.

UMCf-fg = Fine-grained Upper Muddy Creek Formation UMCf-cg = Coarse-grained Upper Muddy Creek Formation

RAMBCLL

Path: H:\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 5-7 - Model Hydraulic Conductivity in Layer 9.mxd

UMCf-cg = Coarse-grained Upper Muddy Creek Formation

Model Hydraulic Conductivity in Layer 10

Path: H:\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 5-8 - Model Hydraulic Conductivity in Layer 10.mxd

UMCf-cg = Coarse-grained Upper Muddy Creek Formation

5-9

1
נMiles

Model Boundary Conditions in Layer 2 Nevada Environmental Response Trust Site Henderson, NV

- Sheet Piles
- Slurry Walls

Horizontal Flow Barrier Boundary

Frenchman Mountain Geologic Fault

Model Boundary Type

Specified Flux Boundary

NERT Property Boundary

Phase 6 Model Extent

EXPLANATION

Horizontal Flow Barrier Boundary

- Slurry Walls
- Frenchman Mountain Geologic Fault

- Specified Flux Boundary
- NERT Property Boundary
- Phase 6 Model Extent

Model Boundary Type

EXPLANATION

Horizontal Flow Barrier Boundary

Frenchman Mountain Geologic Fault

- Specified Flux Boundary
- NERT Property Boundary
- Phase 6 Model Extent

1
נMiles

Model Boundary Conditions in Layer 4 to 10 Nevada Environmental Response Trust Site Henderson, NV $\begin{array}{|c|c|} \hline \textbf{0} & \textbf{0.5} \ \hline \textbf{1} & \textbf{1.5} \$

Model Boundary Type

Path: \\wcevlfps1\ENG\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 6-1 - Model Groundwater Elevation Targets in 2014.aprx

Path: \\wcevlfps1\ENG\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 6-2 - Model Groundwater Elevation Targets in 2018.aprx

Vertical Head Difference Target Location Nevada Environmental Response Trust Site Henderson, NV

Path: H:\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 6-3 - Vertical Head Difference Target Location.mxd

EXPLANATION

- \otimes Transducer Clusters
- **O** Well Cluster
- **O** Well Cluster with Reversed Simulated
Vertical Gradient
	- **NERT Property Boundary**
	- Phase 6 Model Extent

1
נMiles

0 0.5

 $\mathbb N$

1
נMiles

ramboll\projects\NERT\Modeling\Phase_6\figures\ph6_LVW_stream_plot.py

ramboll\projects\NERT\Modeling\Phase_6\figures\ph6_LVW_stream_plot.py

ramboll\projects\NERT\Modeling\Phase_6\figures\ph6_LVW_stream_plot.py

Path: H:\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 6-16 - Capture Zone- Middle.mxd

Path: H:\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 6-17 - Capture Zone- Deep.mxd

Simulated versus Observed Potentiometric Surface Map for Second Quarter 2018 Nevada Environmental Response Trust Site

Notes:

ft amsl = feet above mean sea level

*Ramboll. 2018. Annual Remedial Performance Report for Chromium and Perchlorate, Nevada Environmental Response Trust Site; Henderson, Nevada; July 2017-June 2018. November 9.

.
Miles

 $0.5\,$

EXPLANATION

NERT Off-Site RI Study Area

NERT Property Boundary

Phase 6 Model Boundary

Simulated Heads (ft amsl)

----- Observed Groundwater Elevations (ft amsl)*

City of Henderson Bird Viewing Ponds

N

Path: \\wcevlfps1\ENG\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 8-2 - Model Perchlorate Targets in 2018.aprx

Datum: North American 1983

 N

Nevada Environmental Response Trust Site Henderson, Nevada

Phase 6 Model Extent **Perchlorate Concentration in mg/L**

Note:

mg/L = milligrams per liter

EXPLANATION

Nevada Environmental Response Trust Site Henderson, Nevada

Phase 6 Model Extent **Perchlorate Concentration in mg/L**

Note:

mg/L = milligrams per liter

EXPLANATION

Nevada Environmental Response Trust Site Henderson, Nevada

Phase 6 Model Extent **Perchlorate Concentration in mg/L**

Note:

mg/L = milligrams per liter

EXPLANATION

Nevada Environmental Response Trust Site Henderson, Nevada

Phase 6 Model Extent **Perchlorate Concentration in mg/L**

Note:

mg/L = milligrams per liter

EXPLANATION

Nevada Environmental Response Trust Site Henderson, Nevada

Phase 6 Model Extent **Perchlorate Concentration in mg/L**

Note:

mg/L = milligrams per liter

EXPLANATION

Nevada Environmental Response Trust Site Henderson, Nevada

Phase 6 Model Extent **Perchlorate Concentration in mg/L**

Note:

mg/L = milligrams per liter

EXPLANATION

Nevada Environmental Response Trust Site Henderson, Nevada

Phase 6 Model Extent **Perchlorate Concentration in mg/L**

Note:

mg/L = milligrams per liter

EXPLANATION

Nevada Environmental Response Trust Site Henderson, Nevada

Phase 6 Model Extent **Perchlorate Concentration in mg/L**

Note:

mg/L = milligrams per liter

EXPLANATION

Nevada Environmental Response Trust Site Henderson, Nevada

Phase 6 Model Extent **Perchlorate Concentration in mg/L**

Note:

mg/L = milligrams per liter

EXPLANATION

Nevada Environmental Response Trust Site Henderson, Nevada

Note:

mg/L = milligrams per liter

Perchlorate concentrations in Layer 10 of the model are below 1 mg/L.

EXPLANATION

Phase 6 Model Extent

Perchlorate Concentration in mg/L

Path: \\wcevlfps1\ENG\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Final_MXDs\Figure 9-11- Observed versus Simulated Perchlorate Plume for 2018.mxd

 < 1 mg/L 1-10 10-25 25-100 100-250 250-500

*Ramboll. 2018. Annual Remedial Performance Report for Chromium and Perchlorate, Nevada Environmental Response Trust Site; Henderson, Nevada; July 2017-June 2018. November 9.

EXPLANATION

Off-site NERT RI Study Area

NERT Property Boundary

Phase 6 Model Boundary

Simulated Perchlorate Plume

500-1,000

 \vert >1,000 mg/L

Observed 2018 Plume Contours*

1 mg/L (lines darken as values increase)

Note:

mg/L = milligrams per liter

Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

> **APPENDIX A EVALUATION OF DENSITY DRIVEN FLOW**

A. EVALUATION OF DENSITY-DRIVEN FLOW

The density of a contaminant plume may contribute to the direction of solute transport if dissolved concentrations of contaminants are large enough (Johnson et al. 1989). To investigate the effect of density on flow and transport of perchlorate in the subsurface of the NERT Site, a simple three-dimensional model was constructed to represent the subsurface geology beneath the Site up to a depth of approximately 500 feet (ft). The main purpose of this evaluation was to investigate the potential significance of density-driven flow on vertical migration of the perchlorate plume. The variable density groundwater flow and transport code SEAWAT (Guo and Langevin 2002) was employed and the results were compared with scenarios without density-driven flow.

A.1 Model Setup

A simple three-dimensional MODFLOW model was constructed with 34 layers representing 40 ft of alluvium overlying 450 ft of Upper Muddy Creek formation (UMCf) fine-grained unit (UMCf-fg). The top boundary is the groundwater table. The hydraulic parameters were taken from the Phase 5 groundwater flow model (Ramboll Environ 2016). Hydraulic conductivities of 40 feet per day (ft/d) and 0.72 ft/d were applied to the alluvium and UMCf-fg, respectively. Anisotropy ratios were assumed to be 0.1 for the alluvium and 0.01 for the UMCf-fg. The flow regime was defined by constant head boundary conditions at the south (upgradient) and north (downgradient) boundaries producing a horizontal hydraulic gradient of 0.02 in the domain. An average upward vertical gradient of 0.2 ft/ft was specified as observed in onsite well pairs M-71 and M-162.

The simulation was run for 50 years representing the period when sources of perchlorate to groundwater were active. A constant perchlorate concentration of 50,000 mg/L in the top layer was applied for the entire modeling period. The initial concentration of perchlorate was set to 0 mg/L in the aquifer. The model also incorporates the dual porosity mechanism. An estimated industrial recharge rate of 7.3E-4 ft/d (Ramboll Environ 2016) was applied to the top model boundary. The boundary conditions and model schematic setup is presented in Figure A-1.

A.2 Results and discussion

A total of four scenarios (8 runs in each scenario) were simulated by perturbing source concentration, recharge rate, vertical gradient change, and mass transfer rate between mobile and immobile domains. The recharge rate was perturbed by factors of 1, 10, 50 and 100 with respect to current estimated industrial recharge rate of 7.3E-4 ft/d as benchmark. The source concentration increased up to 200,000 mg/L within the solubility limit of perchlorate. The nominal value of vertical gradient (0.2) was changed between 0 vertical gradient to maximum observed vertical gradient (0.65) at well cluster CMT-201 through CMT-207. The mass transfer rate also increased incrementally from zero, for single porosity model to three orders of magnitude higher than its nominal value (3.0E-5 1/d).

For each scenario, a base case was run with density coupling deactivated that was simulated using MT3D. Three observation points were placed beneath the center of source area (assumed as unit building area) at 40, 70 and 100 ft below the groundwater table representing base of alluvium, shallow UMCf-fg, and the deeper UMCf-fg (Figure A-1). Concentration changes were monitored for each run and compared with respect to base case scenario for each observation point. For each scenario normalized concentrations at each observation point were compared to the base case with no density effects.

The sensitivity parameter runs are presented in Table A-1 with the respective maximum concentration difference between the density and non-density models for each scenario at each observation point at the end of the simulations. The model sensitivity results are also presented in Figure A-2. As expected, the model is most sensitive to the initial perchlorate concentration which in the most extreme case can increase the concentration up to 9% with respect to non-density model. The maximum effect of varying the recharge rate is a 2.5% increase and of varying the mass transfer rate and vertical upward gradient a 2% increase.

A.3 Summary

This simple modeling exercise assumed a high source concentration for a 50-yr loading period. The difference between the model results with and without density-driven flow was small. This indicates that density-driven flow has a negligible effect on vertical migration of perchlorate in areas such as at Unit Building 4 where high total dissolved solid concentrations have been observed. However, this study does not rule out the possible effects of densitydriven flow on localized flow regimes.

LIST OF TABLES

Table A-1 Parameter Sensitivity Analyses for Density Driven Evaluation

LIST OF FIGURES

- Figure A-1 Density Driven Model Schematic
- Figure A-2 Density Driven Model Results

TABLE A-1. SENSITIVITY RUNS PARAMETERS AND SUMMARY

Phase 6 Groundwater Flow and Transport Model

Nevada Environmental Response Trust Site Henderson, Nevada

Notes:

 $1/d = \text{per day}$

ft/d = feet per day

ft/ft = foot per foot

 $g/L =$ grams per liter

Summary of Sensitivity Analysis - Density Effect on Vertical Transport

PROJECT: 1690011200 -034

Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

> **APPENDIX B EVAPOTRANSPIRATION ZONES (2014-2018)**

\\wcevlfps1\ENG\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Appendix B- Evaporation Rates\Fig B-2_ETAreas_2015.mxd

\\wcevlfps1\ENG\LePetomane\NERT\Modeling\Phase 6 Model\Report\Figures\Appendix B- Evaporation Rates\Fig B-3_ETAreas_2016.mxd

-
-

Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

> **APPENDIX C AREAL RECHARGE COMPONENTS (2014 -2018)**

RECHARGE ZONES - 2014 Nevada Environmental Response Trust Site Henderson, Nevada

EXPLANATION

EXPLANATION

EXPLANATION DRAFT Phase 6 Model Extent Undeveloped Residential Area Industrial Area NERT Retention Basins Golf Course Pioneer Detention Basin Bird View Preserve (BVP) Athens Main Drainage Channel TIMET Recharge Trenches Tuscany Sub-drains OSSM Recharge Trenches Pond 13 $W \left(\bigvee_{S}^{N} E \right)$ 0 2,500 5,000 \equiv Feet RAMBCLL **RECHARGE ZONES - 2018** Nevada Environmental Response Trust Site Henderson, Nevada 1690011200-034 Contract Number: Figure Date: 9/16/2019 **C-5** Drafter: Approved: Revised: AS

Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

> **APPENDIX D SUMMARY OF HISTORICAL HYDRAULIC TESTING**

D. INTRODUCTION

This appendix presents the results of numerous aquifer tests and laboratory physical parameter tests conducted within the model domain of the Nevada Environmental Response Trust (Trust) Site located in Henderson, Nevada (the "Site") since 1980. Data from all known tests were used to update the distributions of hydraulic conductivity and storage parameter estimates for lithologic units included in the Phase 6 model (see Section 5.3). A discussion of how testing results were statistically summarized is included in this appendix.

D.1 Aquifer Testing Results Datasets

Results from aquifer testing and laboratory physical parameter testing were compiled from both recent investigations and historic documents. Estimated horizontal hydraulic conductivity, estimated vertical hydraulic conductivity, and estimated storage parameters from these sources are included in Tables D-1, D-2, and D-3, respectively. The data's sources are noted in the tables.

D.1.1 Recent Studies and Investigations

Horizontal hydraulic conductivity values estimated during the following recent studies and investigations were included in the aquifer testing results dataset presented in Table D-1[1:](#page-202-0)

- **Phase 1 RI:** Slug tests were completed at select wells screened in the middle WBZ within the fine-grained UMCf in early 2015.
- **Phase 2 RI:** Select wells were slug tested in Fall 2017 and March 2019 as part of the Phase 2 RI.
- Phase 3 RI: Slug tests were performed in select wells between February and October 2018.
- **Unit 4 and 5 Investigation:** Slug tests were conducted in select wells located near Unit Buildings 4 and 5 in OU-1.
- Pilot and Treatability Studies: Slug, pumping, and specific capacity tests were conducted as part of the Seep Well Field Area Bioremediation Treatability Study, Galleria Drive Bioremediation Treatability Study, Las Vegas Wash Bioremediation Pilot Study, Unit 4 In-Situ Bioremediation Treatability Study, In-Situ Chromium Treatability Study, and the AP Area Treatability Study.

D.1.2 Historic Documents

Since the 1980s, over forty documents that include results from aquifer testing and laboratory physical parameter testing from different areas of the Site have been produced. Estimated horizontal hydraulic conductivity, estimated vertical hydraulic conductivity, and estimated storage parameters from these reports are included in Tables D-1, D-2, and D-3, respectively. Due to inconsistent well/boring names and varied report formats, there is uncertainty that every historic aquifer test result was labeled with the appropriate location and screened lithology. The following quality analysis efforts were taken to minimize error: Location information is included from the January 2019 NDEP All Wells Database (AWDB) unless incongruities with the source document were identified (e.g., an aquifer test was conducted at B-1, but a map within the source document depicts B-1 in a different location than the AWDB location). In these instances, coordinates from the source document were used if available.

ł ¹ Results from aquifer testing conducted by AECOM as part of the Downgradient Investigation were not available for this analysis.

Similarly, the lithologic unit of the well's screened interval was assigned from the AWDB unless the source report contained information about which aquifer unit was tested. The AWDB lithologic unit was overwritten for:

- Results from aquifer tests conducted in the paleochannels or Las Vegas Wash (as estimated in the Phase 5 Model), which were assigned distinct lithologic units since the hydraulic properties in those areas differ from those of the Quaternary Alluvium (Qal); and
- A limited number of locations (H-36, H-43, I-AA, M-17, PC-133), where it was determined that the hydraulic conductivity estimate was more representative of another lithologic unit after the depth to the Upper Muddy Creek formation (UMCf) at the location, the depth of that location's screened interval, and the overall distribution of hydraulic conductivity estimates for each lithologic unit were evaluated.

The variety of aquifer test methods used at the Site by different contractors were categorized as slug, pump, lab, or specific capacity tests. Slug tests include both falling head and rising head tests, as well as piezometer bail tests. Pump tests include constant rate, constant head, recovery, step, and tracer tests. Lab tests include flexible wall permeameter tests in addition to falling and constant head permeability tests.

D.2 Statistical Summary of Hydraulic Conductivity Estimates

The maximum, minimum, and geometric mean hydraulic conductivity for each lithologic unit in the model domain was calculated using the average hydraulic conductivities for each test method for all well locations.

Some aquifer test results were excluded before average hydraulic conductivities at each well location were calculated. Results were excluded in the following instances:

- The reported result was a transmissivity value and no aquifer thickness was available for that test location;
- The lithologic unit in which aquifer testing occurred was unclear, either because location information could not be identified or well/boring logs were not available; or
- The result was reported as an average and the source of the original values was unclear. In these cases, including an average hydraulic conductivity in the calculation of a new average hydraulic conductivity could cause inadvertent weighting of individual test results.

Figure D-1 shows the statistical distribution of average horizontal hydraulic conductivity estimates for each method within each lithologic unit. The statistics for horizontal and vertical hydraulic conductivity are also summarized in Table 4.

LIST OF TABLES

- Table D-1 Summary of Estimated Horizontal Hydraulic Conductivity Table D-2 Summary of Estimated Vertical Hydraulic Conductivity
- Table D-3 Summary of Estimated Storage Parameters

LIST OF FIGURES

Figure D-1 Average Hydraulic Conductivity of Lithologic Units

Notes:

Qal= Quaternary Alluvium

UMCf = Upper Muddy Creek Formation

xMCf = Transitional Muddy Creek Formation

ft/d = feet per day

-- = no data available

¹ Litholologic unit classification for wells was taken from the NDEP All Wells Database or technical report description. Some minor modifications were made as described in Section 1.1.2 of Appendix D.

 2 References for hydraulic properties measured within the Study Area

Data in shaded rows were excluded from statistical summary because they were determined to likely be duplicative of other data, they could not be assigned to a lithologic unit, or they were associated with non-standard site conditions (i.e., a demonstration weir).

Sources:

AECOM. 2011. Revised Upper Muddy Creek Formation Additional Investigation Report, OSSM Joint Groundwater Program, Hederson, Nevada. November 23.

AECOM. 2014. Revised DNAPL Investigation Summary Report Montrose Chemical Corporation of California Environmental Conditions Investigation, Henderson, Nevada. May 15. Prepared for Montrose Chemical Corporation of California.

American Pacific Corporation (AMPAC). 2011a. Semi-Annual Performance (January 1-June 30, 2010) and 2010 Annual Groundwater Monitoring Report. January 12.

AMPAC. 2011b. Remediation System Work Plan for the AMPAC Groundwater Treatment System (AGTS) at Henderson, Nevada. June 2.

Batista, J.R., et al. 2003. The Fate and Transport of Perchlorate in a Contaminated Site in the Las Vegas Valley. March. Prepared for USEPA.

Converse Consultants. 1986. Geotechnical Site Investigation Las Vegas Wash Crossing of the Lass Vegas Valley Lateral Pipeline, Las Vegas, Nevada. June 4. Prepared for Boyle Engineering Corporation.

Converse Consultants. 1999. Preliminary Geotechnical and Geologic Investigation: Industrial Non-Hazardous Disposal Facility, Basic Management Incorporated, Clark County, Nevada. October 27. Prepared for Parsons Engineering Science, Inc.

Converse Consultants. 2002. Geotechnical Data Report: Southern Nevada Water Authority Contract No. 170A, Rainbow Gardens Weir W-3.5 Las Vegas Wash, 1/2 Mile Upstream of Lake Las Vegas, Clark County, Nevada. June 27. Prepared for Boyle Engineering Corporation.

Converse Consultants. 2006a. Geotechnical Data Report: Southern Nevada Water Authority Contract No. 810X, Upper Narrows Weir 7,500 Feet East of Broadbent Boulevard at the Las Vegas Wash, Clark County, Nevada. February 21.

Converse Consultants. 2006b. Geotechnical Data Report: Southern Nevada Water Authority Contract No. 810G, Duck Creek Confluence Weir 1 Mile East of Sam Boyd Stadium at the Las Vegas Wash, Clark County, Nevada. February 22.

Converse Consultants. 2009. Limited Hydrogeologic Investigation: BMI Common Areas (Eastside), Henderson, Nevada. Prepared for Basic Remediation Company. November 25. Crowley Environmental. 2009. Interim Groundwater Capture Evaluation and Vertical Delineation Report, Tronox LLC, Henderson, Nevada. December 23. Prepared for Tronox LLC.

Errol L. Montgomery & Associates, Inc (Montogomery). 2000. Analysis of Rate of Groundwater Movement Based on Results of Tracer and Hydraulic Tests Conducted Between Pittman Lateral and Seep Area, Henderson, Nevada. December 19. Prepared for Kerr-McGee Chemical, LLC.

Geosyntec Consultants (Geosyntec). 2010. Groundwater Flow Model: South of Warm Springs Study Area Henderson, Nevada. March 31. Prepared for American Pacific Corporation.

Geosyntec. 2014. Revised Conceptual Site Model for the Closed Ponds Area and Former Tank Farm Site Assessment Areas, Henderson, Nevada. January 17. Prepared for Montrose Chemical Corporation of California.

Geotechnical & Environmental Services, Inc. (GES) 2003. Geotechnical Investigation: Demonstration Weir W-3.9 Replacement SNWA Contract No. 810-K, Las Vegas Wash Erosion Control Program, Clark County, Nevada. August 8.

GES. 2005a. Geotehcnical Investigation: Homestead Weir W-8.5 SNWA Contract No. 810L, Las Vegas Wash Erosion Control Program, Clark County, Nevada. December 7.

GES. 2005b. Geotehcnical Investigation: Lower Narrows Weir W-8.5 SNWA Contract No. 810M, Las Vegas Wash Erosion Control Program, Clark County, Nevada. December 7.

GES. 2007a. Geotechnical Investigation: Sunrise Mountain Outfall Weir SNWA Contract No. 810W, Las Vegas Wash Erosion Control Program, Clark County, Nevada. February 23.

GES. 2007b. Geotechnical Investigation: Relocated Homestead Weir SNWA Contract No. 810M, Las Vegas Wash Erosion Control Program, Clark County, Nevada. November 1. GES. 2007c. Geotechnical Investigation: Relocated Lower Narrows Weir SNWA Contract No. 810M, Las Vegas Wash Erosion Control Program, Clark County, Nevada. November 1.

Geraghty & Miller, Inc. 1980. Ground-Water Investigation Stauffer Chemical Company, Henderson, Nevada. October.

Hydraulic Conductivity Measurements From Step Discharge Tests in ART Wells, Athens Road Well Field (Nov 2001). Excel file.

Hydraulic Conductivity Measurements from Step Discharge Tests in New Seep Wells, Feb. 2003. Excel File.

K values from ART well pump tests. 2007. Correspondence between Ed Krish of Frontier Net and Brian Giroux of McGinley.

Kerr-McGee Chemical Corporation (KMCC). 1985. Geohydrological Investigation Kerr-McGee Chemical Corporation Hendreson Facility. July. Prepared By Hydrology Services Group Assessment and Remediation Department Kerr-McGee Safety and Environmental Affairs Division.

KMCC. 1998. Preliminary Report on a Hydrogeologic Investigation of Channel-fill Alluvium at the Pittman Lateral, Henderson, Nevada. Prepared By Hydrology Services Group Assessment and Remediation Department Kerr-McGee Safety and Environmental Affairs Division.

Kleinfelder. 2007a. Implementation of the Work Plan for Shallow Well Installation and Associated Soil Sampling, BMI Common Area Eastside, Henderson, Nevada. October 8. Prepared for Basic Remediation Company.

Kleinfelder. 2007b. Implementation of the Revised Aquifer Testing Work Plan: BMI Common Area Eastside, Henderson, Nevada. November 16.

Kleinfelder. 2008. Slug Test Results CAMU Area, Henderson, Nevada. Letter to Basic Remediation Company. January 25.

McGinley & Associates (McGinley). 2003. Las Vegas Wash Initial Perchlorate Modeling Report. October 20. Prepared for Nevada Division of Environmental Protection.

McGinley. 2007. Athens Road Well Field Modeling Report: Near BMI Industrial Complex, Henderson, Nevada. June 30. Prepared for Nevada Division of Environmental Protection. Ninyo & Moore. 2004. Geotechnical Evaluation: SNWA Contract No. 810-R, Landfill Weir (W-5.1), Clark County, Nevada. March 19.

PC-133 Pump Test Analyses, Test Date: December 10-11, 2004. Excel file.

Ramboll Environ. 2013 GWETS Optimization Project Report. Excel file.

Ramboll Environ. 2016. Historical Aquifer Testing Data Compilation. Continuous Optimization Program Hydrogeologic Investigation.

Seep Aquifer Parameters 2007. Info from Tom 8/2/07. Word file.

Stauffer Chemical Company (Stauffer). 1983. 1982 Hydrogeologic Investigation in Support of the Ground-Water Intercept System at Stauffer Chemical Company Henderson, Nevada. March.

Tetra Tech EM Inc. 1998. Draft Environmental Conditions Investigation Report, Titanium Metals Corporation Facility Environmental Conditions Investigation, Henderson, Nevada, Volume 1 of 3. February 20. Submitted to Nevada Division of Environmental Protection.

Tetra Tech. 2016. Final Seep Well Field Area Bioremediation Treatability Study Work Plan, Nevada Environmental Response Trust Site, Henderson, Nevada. September 6. Tetra Tech. 2018a. AP Area Down and Up Flushing Treatability Study Results Report, Nevada Environmental Response Trust Site, Henderson, Nevada. Appendix E Aquifer Testing Results Technical Memorandum. December 21.

Tetra Tech. 2018b. In-Situ Chromium Treatability Study Results Report, Nevada Environmental Response Trust Site, Henderson, Nevada. March 22.

Tetra Tech. 2018c. Unit 4 Source Area In-Situ Bioremediation Treatability Study Work Plan, Revision 1, Nevada Environmental Response Trust Site, Henderson, Nevada. February 5.

Tetra Tech. 2019a. Galleria Drive Bioremediation Treatability Study Work Plan Addendum, Nevada Environmental Response Trust Site, Henderson, Nevada. March 29. Tetra Tech. 2019b. Las Vegas Wash Bioremediation Pilot Study Work Plan Addendum, Nevada Environmental Response Trust Site, Henderson, Nevada. February 15. Tetra Tech. 2019c. Unit 4 and 5 Buildings Investigation Source Area Characterization Report. April.

Titanium Metals Corporation (TIMET). 2007. Conceptual Site Model Titanium Metals Corporation Facility Henderson, Nevada. April 25. Submitted to NDEP.

TIMET. 2008. Tasks I, II, and III, Data Transmittal Report in support of the Groundwater Remedial Alternatives Study. May 16.

TIMET Well Field Pumping Test Results. 2009. Excel file.

Tronox. 2010. Appendix C, Capture Zone Evaluation Report, Henderson, Nevada. December 3.

Wittman G.P. and Carter G.A. 2007. Kleinfelder Letter to Ranajit Sahu Basic Remediation Company. Subject: Slug Test Results BMI Common Area Henderson, Nevada. November 29.

[1] Preliminary results of aquifer testing conducted as part of the NERT Remedial Investigation (RI) Phase 2.

[2] Preliminary results of aquifer testing conducted as part of the NERT RI Phase 3.

TABLE D-2. SUMMARY OF ESTIMATED VERTICAL HYDRAULIC CONDUCTIVITY

Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

TABLE D-2. SUMMARY OF ESTIMATED VERTICAL HYDRAULIC CONDUCTIVITY

Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

TABLE D-2. SUMMARY OF ESTIMATED VERTICAL HYDRAULIC CONDUCTIVITY

Phase 6 Groundwater Flow and Transport Model Nevada Environmental Response Trust Site Henderson, Nevada

TABLE D-2. SUMMARY OF ESTIMATED VERTICAL HYDRAULIC CONDUCTIVITY Phase 6 Groundwater Flow and Transport Model

Nevada Environmental Response Trust Site

Henderson, Nevada

Notes:

Qal= Quaternary Alluvium

UMCf = Upper Muddy Creek Formation

xMCf = Transitional Muddy Creek Formation

 $ft/d = feet$ per day

 1 Litholologic unit classification for wells was taken from the NDEP All Wells Database or technical report description. Some minor modifications were made as described in Section 1.1.2 of Appendix D.

 2 References for hydraulic properties measured within the Study Area

Sources:

Geraghty & Miller, Inc. 1980 Geraghty & Miller, Inc. 1980. Ground-Water Investigation Stauffer Chemical Company, Henderson, Nevada. October. Northgate 2009 Northgate. 2009. Interim Groundwater Capture Evaluation and Vertical Delineation Report; Tronox LLC; Henderson, Nevada. December 23.

Northgate 2010 Northgate. 2010. Capture Zone Evaluation Report, Henderson, Nevada. December 10. ENVIRON resubmitted the NERT Site groundwater model on February 21, 2013. NDEP approved April 4, 2013.

TIMET. 2008. Tasks I, II, and III, Data Transmittal Report in support of the Groundwater Remedial Alternatives Study. May 16.

[1] These data are preliminary results of laboratory physical testing conducted as part of the NERT Remedial Investigation (RI).

Notes:

Qal= Quaternary Alluvium UMCf = Upper Muddy Creek Formation

UMCf-cg = Upper Muddy Creek Formation, coarse-grained

xMCf = Transitional Muddy Creek Formation

 $ft = feet$

-- = not applicable or unavailable

¹ Litholologic unit classification for wells was taken from the NDEP All Wells Database or technical report description. Some minor modifications were made as described in Section 1.1.2 of Appendix D.

² References for hydraulic properties measured within the Study Area

Sources:

American Pacific Corporation (AMPAC). 2011a. Semi-Annual Performance (January 1-June 30, 2010) and 2010 Annual Groundwater Monitoring Report. January 12.

AMPAC. 2011b. Remediation System Work Plan for the AMPAC Groundwater Treatment System (AGTS) at Henderson, Nevada. June 2.

Batista, J.R., et al. 2003. The Fate and Transport of Perchlorate in a Contaminated Site in the Las Vegas Valley. March. Prepared for USEPA.

Converse Consultants. 1986. Geotechnical Site Investigation Las Vegas Wash Crossing of the Lass Vegas Valley Lateral Pipeline, Las Vegas, Nevada. June 4. Prepared for Boyle Engineering Corporation.

Converse Consultants. 2002. Geotechnical Data Report: Southern Nevada Water Authority Contract No. 170A, Rainbow Gardens Weir W-3.5 Las Vegas Wash, 1/2 Mile Upstream of Lake Las Vegas, Clark County, Nevada. June 27. Prepared for Boyle Engineering Corporation.

Errol L. Montgomery & Associates, Inc (Montgomery). 2000. Analysis of Rate of Groundwater Movement Based on Results of Tracer and Hydraulic Tests Conducted Between Pittman Lateral and Seep Area, Henderson, Nevada. December 19. Prepared for Kerr-McGee Chemical, LLC.

Geraghty & Miller, Inc. 1980. Ground-Water Investigation Stauffer Chemical Company, Henderson, Nevada. October.

Kerr-McGee Chemical Corporation (KMCC). 1985. Geohydrological Investigation Kerr-McGee Chemical Corporation Hendreson Facility. July. Prepared By Hydrology Services Group Assessment and Remediation Department Kerr-McGee Safety and Environmental Affairs Division.

KMCC. 1998. Preliminary Report on a Hydrogeologic Investigation of Channel-fill Alluvium at the Pittman Lateral, Henderson, Nevada. Prepared By Hydrology Services Group Assessment and Remediation Department Kerr-McGee Safety and Environmental Affairs Division.

Kleinfelder. 2007. Implementation of the Revised Aquifer Testing Work Plan: BMI Common Area Eastside, Henderson, Nevada. November 16.

McGinley & Associates (McGinley). 2003. Las Vegas Wash Initial Perchlorate Modeling Report. October 20. Prepared for Nevada Division of Environmental Protection.

Ramboll Environ. 2013 GWETS Optimization Project Report. Excel file.

Stauffer Chemical Company (Stauffer). 1983. 1982 Hydrogeologic Investigation in Support of the Ground-Water Intercept System at Stauffer Chemical Company Henderson, Nevada. March.

Tetra Tech EM Inc. 1998. Draft Environmental Conditions Investigation Report, Titanium Metals Corporation Facility Environmental Conditions Investigation, Henderson, Nevada, Volume 1 of 3. February 20. Submitted to Nevada Division of Environmental Protection.

TIMET. 2009. Design Data Gap Investigation. June 12.

APPENDIX E OBSERVED GROUNDWATER ELEVATIONS AT CALIBRATION TARGETS (2014-2018) (PROVIDED ELECTRONICALLY)

APPENDIX F OBSERVED HEAD DIFFERENCE AT CALIBRATION TARGETS (2014-2018) (PROVIDED ELECTRONICALLY)

APPENDIX G OBSERVED VERSUS SIMULATED GROUNDWATER ELEVATIONS AT CALIBRATION TARGETS (PROVIDED ELECTRONICALLY)

APPENDIX H OBSERVED PERCHLORATE CONCENTRATIONS AT CALIBRATION TARGETS (2014–2018) (PROVIDED ELECTRONICALLY)

APPENDIX I OBSERVED VERSUS SIMULATED PERCHLORATE CONCENTRATIONS AT CALIBRATION TARGETS (PROVIDED ELECTRONICALLY)

> **APPENDIX J MODEL FILES (PROVIDED ELECTRONICALLY)**